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Heterogeneous Atmospheric Chemistry, Ambient Measurements, and Model Calculations of N_2O_5 : A Review

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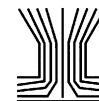
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Heterogeneous Atmospheric Chemistry, Ambient Measurements, and Model Calculations of N_2O_5 : A Review

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For several decades, dinitrogen pentoxide (N_2O_5) has been recognized as an important reactive intermediate in the atmospheric chemistry of nitrogen oxides and nitrate aerosol, especially during nighttime. However, due to the lack of ambient observations of N_2O_5 , the nocturnal nitrogen oxide chemistry could not be quantified until recent years. The objective of the present article is to assess the current state-of-the-art knowledge of N_2O_5 dynamics within the tropospheric aerosol. An up-to-date summary of N_2O_5 chemistry and major loss mechanisms are provided. Furthermore, techniques for measuring ambient N_2O_5 and an overview of typical N_2O_5 levels in the troposphere are described. In addition, model representations of N_2O_5 chemistry are reviewed along with key features of N_2O_5 vertical profiles based on numerical simulations. Lastly, the article provides the outstanding uncertainties and needs for further research into the atmospheric chemistry of N_2O_5 . These include the need for better characterization of N_2O_5 heterogeneous

uptake under temperature conditions characteristic of mid- and high-latitude winter seasons; greater understanding of the influence of individual aerosol components on N_2O_5 uptake and representation of these components in atmospheric models; and comprehensive descriptions of nighttime vertical profiles of N_2O_5 and related pollutants.

1. INTRODUCTION

Dinitrogen pentoxide (N_2O_5) is an important reactive intermediate in the atmospheric chemistry of nitrogen oxides. It was first recognized as a limiting factor in the destruction of ozone (O_3) by NO and NO_2 in the lower stratosphere (Johnston 1971; Crutzen 1971). In the troposphere, N_2O_5 has long been understood to be a sink for NO_x ($\text{NO}_x = \text{NO} + \text{NO}_2$) through its reaction with water (either in the gas or particle phase) to produce HNO_3 , which has a direct effect on tropospheric ozone production (Morris and Niki 1973; Mozurkewich and Calvert 1988) and particulate matter (PM) formation (Russell and Cass 1986). Although the formation and loss of N_2O_5 is generalized in kinetic mechanisms for photochemical smog (Hecht and Seinfeld 1972; Atkinson et al. 2004), its significance has been open to question for several decades (Calvert 1976; Brown et al. 2006b).

The lifetime of NO_x , a key player in atmospheric chemistry, depends critically on the overnight fate of N_2O_5 , which may either serve as a reservoir that allows for reversible storage and transport or as a sink that leads to efficient overnight removal of NO_x (Dentener and Crutzen 1993). During the night, N_2O_5 is in rapid equilibrium with the nitrate radical, NO_3 , one of the main nocturnal atmospheric oxidants (Wayne et al. 1991).

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A determining factor of NO_3 availability, and thus nighttime oxidative capacity, is the hydrolysis of N_2O_5 , which, depending on its efficiency, may consume a large fraction of the NO_3 produced at night. The gas-phase reaction of N_2O_5 with H_2O is slow and relatively constant, while the heterogeneous reaction is more variable and can be quite rapid on particle surfaces (Mozurkewich and Calvert 1988; Li et al. 1993; Patris et al. 2007). The overnight removal of NO_3 and N_2O_5 directly impacts the production of oxidants such as hydroxyl radical (OH) and ozone at dawn. The resulting impacts on regional O_3 are complex. Due to the nonlinear relationship between NO_x concentration and ozone production, heterogeneous hydrolysis of N_2O_5 is predicted to decrease ozone under low- NO_x conditions and increase ozone in high- NO_x regions (Riemer et al. 2003).

Several air quality modeling studies have demonstrated the influence of tropospheric N_2O_5 on oxidant levels at the urban (Russell et al. 1985; Lei et al. 2004), regional (Riemer et al. 2003; McLaren et al. 2004), and global scales (Dentener and Crutzen 1993; Tie et al. 2001, 2003; Evans and Jacob 2005; Feng and Penner 2007). The tropospheric aerosol budget is also directly influenced by N_2O_5 because its main hydrolysis product, nitric acid, partitions favorably to the aerosol phase at low temperatures or in environments with excess ammonia (Stelson and Seinfeld 1982; Russell and Cass 1986). Direct N_2O_5 uptake by particles and fog droplets is also an important source of dissolved nitrate (Lillis et al. 1999). Particle-phase nitrate dynamics is still a major area of uncertainty in atmospheric aerosol prediction today (Yu et al. 2005) and has significant impacts on regional air quality and climate (Liao and Seinfeld 2005; Feng and Penner 2007; Bauer et al. 2007).

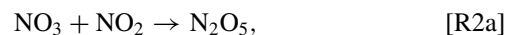
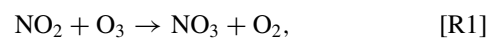
Despite its importance, many aspects of tropospheric N_2O_5 chemistry are still not well known. The rate of N_2O_5 uptake by particles, which depends on aerosol composition and meteorological conditions (relative humidity (RH) and temperature), remains highly uncertain (Brown et al. 2006b; Bertram et al. 2009). Only recently, the heterogeneous chemistry of N_2O_5 on chloride-containing aerosol, known from laboratory studies for two decades (Finlayson-Pitts et al. 1989a), has been shown to efficiently release atomic chlorine radicals to the atmosphere (Osthoff et al. 2008). The impacts of this chemistry on regional and global scales are still unknown. In addition, little is known about the vertical distribution of N_2O_5 in the nocturnal atmosphere. Many of these questions can be answered by novel atmospheric measurement techniques that have emerged only recently.

We provide an overview of the current knowledge of N_2O_5 chemistry in the troposphere and identify the most pressing research needs. After summarizing the known N_2O_5 chemistry and the dominant loss pathways in Section 2, Section 3 describes the ambient N_2O_5 measurement techniques that are currently available. Section 4 presents a discussion of the key observations of ambient N_2O_5 , including typical concentrations and vertical distributions during nighttime. Section 5 outlines the representation of N_2O_5 chemistry in atmospheric models, tro-

sospheric N_2O_5 vertical gradients based on simulation results, and related challenges facing the modeling community. Finally, Section 6 presents our conclusions and identifies areas where further research is needed most urgently.

2. ATMOSPHERIC N_2O_5 CHEMISTRY

N_2O_5 is formed from the gas-phase oxidation of NO_2 by O_3 , followed by the reaction of NO_2 with NO_3 (R1 and R2a):



N_2O_5 is thermally unstable, so it dissociates back to NO_3 and NO_2 [Reaction (R2b)]. The equilibrium between NO_2 , NO_3 , and N_2O_5 is given by

$$[\text{N}_2\text{O}_5] = K_{\text{eq}}[\text{NO}_2][\text{NO}_3], \quad [1]$$

where K_{eq} is a temperature-dependent equilibrium constant ($\text{cm}^3 \text{ molecule}^{-1}$). Cool weather promotes a shift in the equilibrium of Reaction (R2) toward N_2O_5 , as shown in Figure 1. Whereas N_2O_5 and NO_3 are in roughly equal proportions at 1 ppbv NO_2 and 295 K, the ratio of N_2O_5 to NO_3 is approximately 10 at 278 K. In addition, N_2O_5 typically exceeds NO_3 under polluted conditions (e.g., 10 ppbv NO_2).

Two different loss pathways have to be considered for N_2O_5 : direct losses, in which N_2O_5 undergoes hydrolysis reactions, and

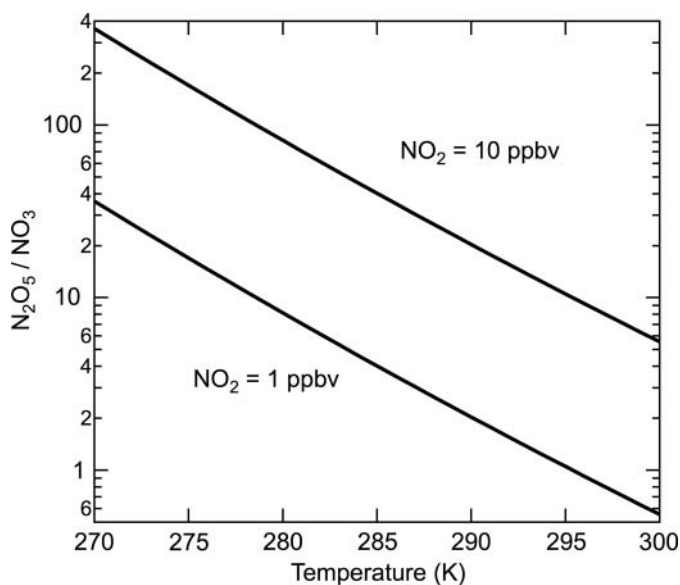


FIG. 1. The $\text{N}_2\text{O}_5/\text{NO}_3$ ratio as a function of temperature using $K_{\text{eq}} = A \cdot \exp(B/T)$, in which T is the temperature in Kelvin, $A = 2.7 \times 10^{-27} \text{ cm}^3 \text{ molecule}^{-1}$, and $B = 11,000$ (Sander et al. 2006). Other recommendations for K_{eq} are in close agreement with this expression (Osthoff et al. 2007).

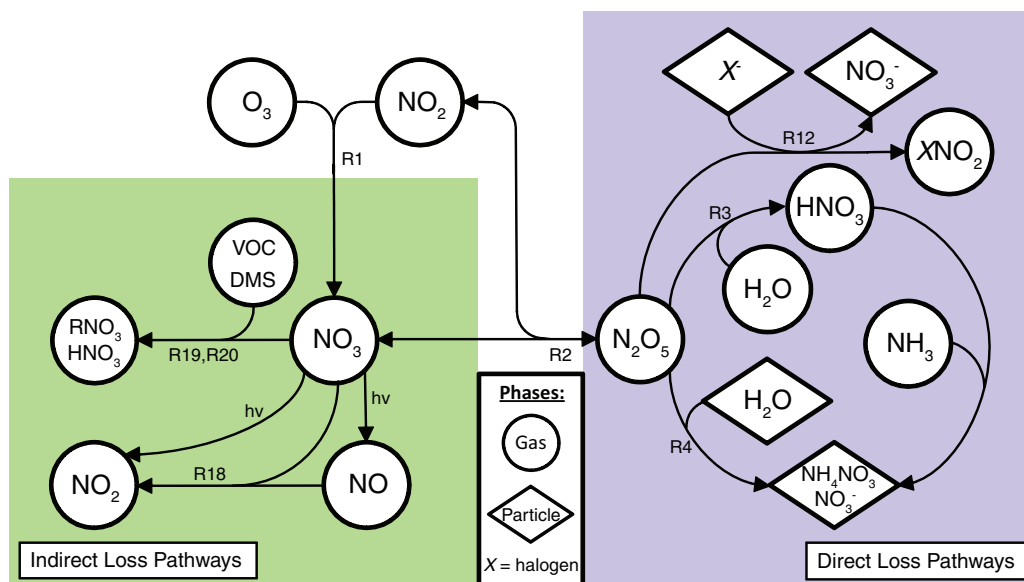
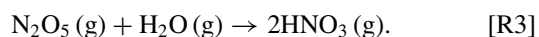


FIG. 2. Chemical mechanism for N₂O₅ formation and loss. X indicates halogen atoms such as chlorine and bromine that are commonly found in sea-salt aerosol. Circles denote gas-phase species, and diamonds denote species in particle phase.

indirect pathways, in which a loss of NO₃ or NO₂ propagates through thermal equilibrium and thus leads to an N₂O₅ decrease. The relative importance of these two pathways is determined by the temperature-dependent equilibrium [Equation (1)] and the presence of various reactants of NO₃, NO₂, and N₂O₅ (including uptake on surfaces). A schematic of the major loss pathways is shown in Figure 2.

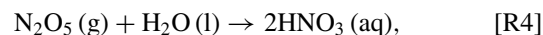
2.1. Direct Loss Pathways

The direct chemical loss of atmospheric N₂O₅ proceeds through its hydrolysis reaction (Morris and Niki 1973),



It is clear today that this reaction is too slow as a second-order process in the gas phase to have significant impact on the lifetime of N₂O₅ and NO_x (Detener and Crutzen 1993). Wahner and coworkers also reported a moderate contribution of Reaction (R3) to the overall hydrolysis rate but suggested that the third-order reaction, N₂O₅(g) + 2H₂O(g), may be important (Mentel et al. 1996; Wahner et al. 1998a). While Voegelé et al. (2003) provide some theoretical insights into the reaction of N₂O₅ with water, further studies of this reaction are still warranted in order to elucidate the magnitude of its potential impact (Section 4.3). Due to the lack of agreement between data obtained from laboratory and field measurement studies, IUPAC (2010) recommend an upper limit of 1 × 10⁻²² cm³ molecule⁻¹ s⁻¹ for the bimolecular process only, until further insights are obtained.

The heterogeneous hydrolysis of N₂O₅ on and within aerosol particles, fog, or cloud droplets,



has been shown to be much faster than homogeneous hydrolysis under typical tropospheric conditions and is now believed to be the dominant pathway for direct N₂O₅ removal (Russell et al. 1985; Dentener and Crutzen 1993; Hanway and Tao 1998). This reaction can be modeled as a pseudo-first-order process (Heikes and Thompson 1983; Chang et al. 1987):

$$\left. \frac{d[\text{N}_2\text{O}_5]}{dt} \right|_{\text{het}} = -k_{\text{N}_2\text{O}_5}[\text{N}_2\text{O}_5], \quad [2]$$

where dt denotes the time increment and $k_{\text{N}_2\text{O}_5}$ represents the rate constant for the heterogeneous surface reaction (Schwartz 1986),

$$k_{\text{N}_2\text{O}_5} = \left(\frac{r}{D_g} + \frac{4}{c_{\text{N}_2\text{O}_5} \gamma_{\text{N}_2\text{O}_5}} \right)^{-1} S. \quad [3]$$

Here, r is the mean particle radius, D_g is the gas-phase diffusion coefficient for N₂O₅, $c_{\text{N}_2\text{O}_5}$ is the mean molecular velocity of N₂O₅, S is the aerosol surface area density, and $\gamma_{\text{N}_2\text{O}_5}$ is the reaction probability. Dentener and Crutzen (1993) showed that when $\gamma_{\text{N}_2\text{O}_5} < 0.1$, the dominant term in Equation (3) is $4/(c_{\text{N}_2\text{O}_5} \gamma_{\text{N}_2\text{O}_5})$, so the reaction rate constant simplifies to (Riemer et al. 2003):

$$k_{\text{N}_2\text{O}_5} = \frac{1}{4} c_{\text{N}_2\text{O}_5} \cdot \gamma_{\text{N}_2\text{O}_5} \cdot S. \quad [4]$$

The reaction probability, $\gamma_{\text{N}_2\text{O}_5}$, is the fraction of collisions between gaseous N_2O_5 molecules and particle surfaces that result in a loss of N_2O_5 . In other words, it is the likelihood of N_2O_5 uptake on particles. This parameter has been measured under controlled laboratory conditions on three major types of substrates: ice surfaces, water droplets, and aerosol surfaces. The studies on ice surfaces are of critical importance at high latitudes and in the stratosphere (Tolbert et al. 1988; Horn et al. 1994; Apodaca et al. 2008), whereas the studies on water droplets are thought to be important under foggy conditions (Wood et al. 2005; Sommariva et al. 2009) and in clouds. Laboratory measurements on those substrates are summarized by Sander et al. (2006) and are beyond the scope of this review.

Laboratory measurements on aerosol surfaces commonly found in the troposphere show that $\gamma_{\text{N}_2\text{O}_5}$ is affected by particle composition, RH, and, to a lesser extent, temperature. Most of the early laboratory studies focused on inorganic particle compositions. Davis et al. (2008) reviewed data from seven laboratory studies (Mozurkewich and Calvert 1988; Hu and Abbatt 1997; Folkers 2001; Kane et al. 2001; Folkers et al. 2003; Hallquist et al. 2003; Badger et al. 2006), in which $\gamma_{\text{N}_2\text{O}_5}$ was measured on ammoniated sulfate and nitrate particles in both crystalline- and aqueous-phase states. Values of $\gamma_{\text{N}_2\text{O}_5}$ on the aqueous particles range from 0.001 at 29% RH to 0.086 at 76% RH, whereas the values on crystalline particles fall between 0.0003 and 0.012. More recent measurements reported by Bertram and Thornton (2009) and Griffiths and Cox (2009) also fall within these ranges. Heterogeneous uptake of N_2O_5 on sodium-containing particles has also received significant attention because sea-salt surfaces can be a sink for nitrogen oxides and a source of halogen radicals [Reaction (R12)]. Values of $\gamma_{\text{N}_2\text{O}_5}$ on aqueous NaCl and natural sea-salt particles range from 0.006 at 30% RH to 0.04 at 77% RH (Behnke et al. 1997; Schweitzer et al. 1998; Stewart et al. 2004; Thornton and Abbatt 2005; McNeill et al. 2006). On aqueous NaHSO_4 and Na_2SO_4 particles, $\gamma_{\text{N}_2\text{O}_5}$ lies at the upper end of that range, between 0.018 and 0.04 (Mentel et al. 1999). In contrast, $\gamma_{\text{N}_2\text{O}_5}$ on aqueous NaNO_3 particles is more than an order of magnitude lower, falling between 0.0003 and 0.003 (Wahner et al. 1998b; Hallquist et al. 2003). Mentel et al. (1999) proposed a mechanism to explain the suppression of $\gamma_{\text{N}_2\text{O}_5}$ by nitrate, which has been extended recently by Griffiths et al. (2009).

Several attempts have been made to unify these laboratory observations in a single chemical mechanism. Mozurkewich and Calvert (1988) proposed an N_2O_5 reaction mechanism, which is consistent with the sensitivities of $\gamma_{\text{N}_2\text{O}_5}$ to RH and aerosol sulfate: N_2O_5 dissolves in water, dissociates to NO_2^+ and NO_3^- , both of which will react with water and H^+ to form HNO_3 .

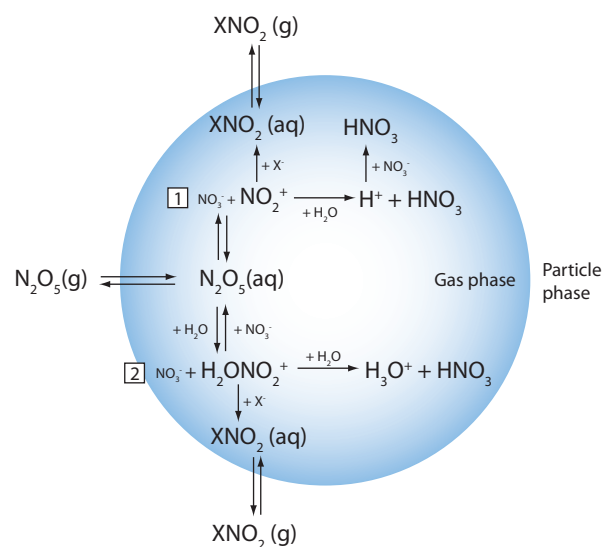
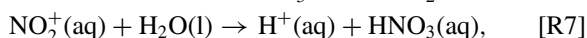
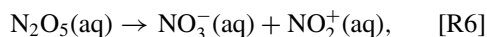
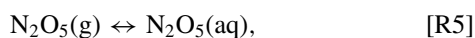
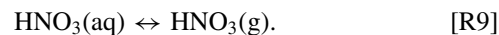
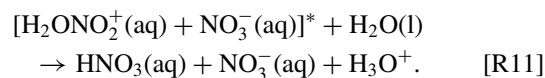


FIG. 3. Schematic depiction of the heterogeneous reaction of N_2O_5 with aqueous particle proposed by Griffiths et al. (2009) in summary of previous studies and Bertram and Thornton (2009). X denotes a halogen species such as chlorine, bromine, and iodine.



This “ionic hydrolysis mechanism” was also supported by other authors, e.g., Wahner et al. (1998b), Mentel et al. (1999), Hallquist et al. (2003), and Griffiths et al. (2009). Wahner et al. (1998) explained the lower $\gamma_{\text{N}_2\text{O}_5}$ for nitrate-containing aerosol by the fact that the recombination reaction $\text{NO}_2^+ + \text{NO}_3^-$ is favored when large nitrate concentrations are present. Thornton et al. (2003) introduced a modified mechanism for dilute acidic or neutral nonhalide aerosol (Figure 3), which is also consistent with observations of the inhibition of N_2O_5 uptake by aerosol nitrate and the formation of halogen nitrites from aqueous sodium halide aerosol [see (R12) below and related discussion]. In this mechanism, Reactions (R6–R8) of the Mozurkewich mechanisms are replaced by

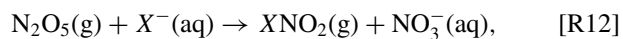


This mechanism involves the formation of the hydrated NO_2^+ intermediate, H_2ONO_2^+ , instead of NO_2^+ . However, Thornton et al. (2003) and Bertram and Thornton (2009) pointed out that H_2ONO_2^+ has thus far never been directly observed in the systems discussed here. On the basis of their results, Bertram and Thornton (2009) developed a parameterization for $\gamma_{\text{N}_2\text{O}_5}$ as a function of the molar ratios of $\text{H}_2\text{O}(\text{l})$ to NO_3^- and Cl^- to NO_3^-

for the use in regional and global models (see Section 5.1 for a discussion of model parameterizations of N₂O₅ hydrolysis).

Mozurkewich and Calvert (1988) argued that the diffusion of N₂O₅ into the drop [Reaction (R5)] was expected to be much slower compared to its dissociation [Reaction (R6)] and therefore concluded that the reaction takes place almost entirely in a thin layer near the surface. Hence, the loss of N₂O₅ should be proportional to the aerosol surface area. Thornton et al. (2003) revisited the question of surface versus volume reaction with inconclusive results. According to Mozurkewich and Calvert (1988), the mechanisms (R6)–(R9) are consistent with reaction probabilities becoming smaller at low humidities because under these conditions only a thin layer of liquid may be available. They argued that when this layer is thin enough, N₂O₅ can diffuse uniformly throughout and hence saturate it. As a result, the rate of reevaporation becomes faster. Similarly, the decrease of $\gamma_{\text{N}_2\text{O}_5}$ with increasing temperature is also consistent with the increase in the rate of evaporation compared to the rate of dissociation.

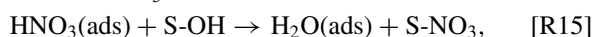
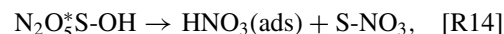
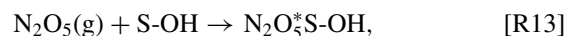
The mechanism by Bertram and Thornton (2009) also explains the formation of nitryl halide molecules, which was previously observed by Finlayson-Pitts et al. (1989a, b) and Barnes et al. (1991). This reaction is often summarized by



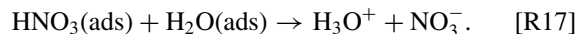
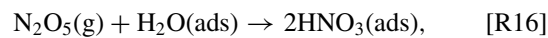
where X denotes a halogen species such as chlorine, bromine, and iodine. In the case of chlorine, the Bertram and Thornton mechanism and Reaction (R12) indicate that the yield of ClNO₂ depends on the Cl[−] concentration within the particle, but observed mixing ratios of ClNO₂ in the marine boundary layer as well as mid-continental areas are much higher than predictions made based on ambient particulate chlorine level (Osthoff et al. 2008; Thornton et al. 2010). Another proposed mechanism for the ambient production of ClNO₂ is the reaction between gaseous HCl and the autoionized intermediate of N₂O₅, NO₂⁺NO₃[−], on water-containing surfaces (Raff et al. 2009). The product distribution from Reaction (R12) as a function of Cl[−] concentration based on field measurements has been investigated in detail and is discussed further in Section 4.4. The subsequent photolysis of nitryl halides produces halogen radicals (Leu et al. 1995), which can increase the oxidative capacity of the atmosphere, thereby impacting ozone and other gaseous pollutants (Simon et al. 2009). When Reaction (R12) takes place on highly acidic particles, the reaction of N₂O₅ with chloride may even produce Cl₂ (Roberts et al. 2008, 2009), which photolyzes rapidly to form two atomic chlorine radicals.

Several studies investigated the uptake of N₂O₅ on mineral dust particles (Seisel et al. 2005; Karagulian et al. 2006; Wagner et al. 2008, 2009). On these substrates, $\gamma_{\text{N}_2\text{O}_5}$ ranges from 0.005 to 0.3. Seisel et al. (2005) proposed two reaction mechanisms of N₂O₅ on dust. The first reaction proceeds via the reaction of gas-phase N₂O₅ with a surface OH group to form surface nitrate and nitric acid. HNO₃(ads) is assumed to readily react

via Reaction (R15) forming another surface nitrate:



and S is surface site. A second pathway proceeds via the hydrolysis of N₂O₅, most likely in the interlamellar water of clay minerals in mineral dust (Seisel et al. 2005):



Sander et al. (2006) summarized the available measurements of $\gamma_{\text{N}_2\text{O}_5}$ on soot particles, which vary broadly from 0.0002 to 0.02 at ambient RH. Results from a more recent study on decane flame soot fall within the 0.005–0.03 range (Karagulian and Rossi 2007).

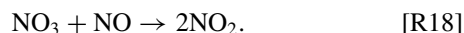
The NO₂⁺-NO₃[−] intermediate is also recognized as an important source of nitration for organic species in the condensed phase, such as the occurrence of nitrophenols (Harrison et al. 2005). Laboratory measurements of $\gamma_{\text{N}_2\text{O}_5}$ on organic particles fall into three categories: pure organic particles, mixed organic/inorganic particles, and inorganic particles coated with organic material. The data on pure organic substrates are limited to five studies (Thornton et al. 2003; Badger et al. 2006; Gross and Bertram 2008; Griffiths et al. 2009; Gross et al. 2009). Most of these studies reported $\gamma_{\text{N}_2\text{O}_5}$ values at least an order of magnitude lower than those measured on aqueous inorganic salts (e.g., 6×10^{-5} on liquid-phase oleic acid and $4\text{--}8 \times 10^{-4}$ on liquid-phase glycerol). The highest reported values are 0.017–0.03 on malonic acid at 50%–70% RH (Thornton et al. 2003; Griffiths et al. 2009). As in the case of inorganic substrates, several studies have reported $\gamma_{\text{N}_2\text{O}_5}$ values on solid-phase organics to be much smaller than those on liquid-phase organics (Thornton et al. 2003; Griffiths et al. 2009; Gross et al. 2009). In addition, uptake on solid polycyclic aromatic hydrocarbons was too slow to be measured (Gross and Bertram 2008). Two laboratory studies have explored $\gamma_{\text{N}_2\text{O}_5}$ on mixed organic/inorganic particles. When humic acid is mixed with increasing amounts of ammonium sulfate ranging from 60% to 94% of the particle mass, $\gamma_{\text{N}_2\text{O}_5}$ increases from 0.0003 to 0.0083 (Badger et al. 2006). Griffiths et al. (2009) found that the addition of dicarboxylic acids to ammonium bisulfate has a small effect on $\gamma_{\text{N}_2\text{O}_5}$, but their effect on uptake by ammonium sulfate particles is more pronounced; small amounts of organic acid enhance $\gamma_{\text{N}_2\text{O}_5}$, while larger amounts inhibit. Numerous laboratory investigations into the effect of organic coatings on $\gamma_{\text{N}_2\text{O}_5}$ have been reported in recent years (Folkers et al. 2003; Thornton and Abbatt 2005; Anttila et al. 2006; McNeill et al. 2006; Park et al. 2007; Cosman et al. 2008; Cosman and Bertram 2008). Overall, these studies indicate that organic coatings can decrease N₂O₅ uptake coefficients by an order of magnitude or

more. Various explanations for the reduction of $\gamma_{\text{N}_2\text{O}_5}$ in the presence of organic films have been proposed. Folkers et al. (2003) concluded that the main cause for the decrease of $\gamma_{\text{N}_2\text{O}_5}$ in the presence of multilayer organic films is a smaller Henry's law constant or mass accommodation coefficient for the films, leading to a slower initial solvation of N_2O_5 in the droplets. Thornton and Abbatt (2005), working with a coating thickness of only a monolayer, proposed a reduction in the accommodation coefficient and the reaction rate at the interface as a possible explanation. Anttila et al. (2006), on the other hand, attributed the decrease of $\gamma_{\text{N}_2\text{O}_5}$ to the fact that the reaction rate constant of N_2O_5 in the coating is decreased by 3–5 orders of magnitude and that the product of the solubility of N_2O_5 and its diffusion coefficient in the coating are reduced by more than an order of magnitude compared to the corresponding value for the aqueous phase.

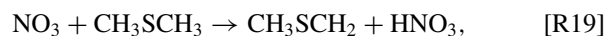
With the recent development of in situ techniques for N_2O_5 measurement (Sections 3.2–3.5), it has become possible to determine $\gamma_{\text{N}_2\text{O}_5}$ on atmospheric aerosol particles, which exhibit more complex chemical composition than those studied in the laboratory setting (Section 4.3). In addition, efforts have been made to incorporate the growing body of $\gamma_{\text{N}_2\text{O}_5}$ measurements into a unified model parameterization (Section 5.1).

2.2. Indirect Loss Pathways

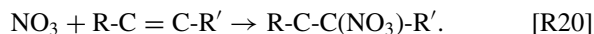
As noted above, indirect loss of N_2O_5 can proceed through the reactive losses of NO_3 . One of the dominant NO_3 loss pathways during the day and at the surface at night is the reaction with NO (Wayne et al. 1991; Asaf et al. 2010):



Together with the daytime photolysis of NO_3 (Johnston et al. 1996), Reaction (R18) leads to low daytime concentrations of NO_3 and N_2O_5 . However, under polluted urban conditions, small but nonnegligible concentrations of gas-phase NO_3 and N_2O_5 can be sustained in the daytime (Geyer et al. 2003; Brown et al. 2005; Osthoff et al. 2006). Other major NO_3 loss pathways include reaction with dimethyl sulfide (DMS) and hydrocarbons (Allan et al. 2000), as well as heterogeneous uptake (Moise et al. 2002; Karagulian and Rossi 2007; Gross et al. 2009). The NO_3 reaction with DMS,



is most important in marine environments (Allan et al. 1999). The reaction with hydrocarbons may also initiate with the abstraction of hydrogen, but with species such as isoprene and terpenes in particular, the reaction is predominately via addition to the carbon–carbon double bond,



Heterogeneous NO_3 reaction rates with aerosols vary with particle compositions and meteorological conditions, analogues to the previously summarized heterogeneous uptake of N_2O_5 . While the uptake coefficient of NO_3 on mineral dust is similar in magnitude to the values for N_2O_5 , uptake on organics favors NO_3 over N_2O_5 (Tang et al. 2010). Karagulian and Rossi (2005) reported γ_{NO_3} values on mineral dusts ranging between 0.034 and 0.12. For soot particles, measured peak γ_{NO_3} values vary from 0.0003 to 0.05 (Saathoff et al. 2001; Mak et al. 2007; Karagulian and Rossi 2007). For organic-containing species, 10^{-3} was proposed to be the upper limit of the NO_3 reactive uptake coefficient (McNeill et al. 2007) and 0.015 for liquid and frozen organics (Moise et al. 2002). However, more recent measurements demonstrated much larger uptake coefficient values on organic aerosols. Reported γ_{NO_3} values include 0.059–0.79 for polycyclic aromatic hydrocarbons (Gross and Bertram 2008), 0.034 for terminal alkene monolayer (Gross and Bertram 2009), and 0.076–0.62 for liquid alkenoic acids (Gross et al. 2009).

2.3. Overall Nighttime NO_x Loss

Whether through direct or indirect pathways, the loss of N_2O_5 is a controlling factor in the nighttime lifetime of NO_x . Although Reaction (R1) is slow, it is fast enough that it converts a large fraction of the NO_x available at sunset within a given air mass into NO_3 . If N_2O_5 hydrolysis is rapid (i.e., first-order rate coefficient larger than $2.5 \times 10^{-4} \text{ s}^{-1}$, corresponding to an uptake coefficient, $\gamma = 0.02$, and an aerosol surface area $>200 \mu\text{m}^2 \text{ cm}^{-3}$), it effectively doubles the rate coefficient for this conversion since Reaction (R1) becomes rate limiting and each NO_3 radical produced goes on to react with a second NO_2 via Reaction (R2a) with very little dissociation [Reaction (R2b)]. For example, during summertime conditions (i.e., 298 K) with a nominal (constant) background of 50 ppbv O_3 , the reaction sequence would convert 94% of NO_x to HNO_3 (or $\text{HNO}_3 + \text{XNO}_2$) in a 10-h night when estimated with a model; in winter (i.e., 273 K), Reaction (R1) is much slower, but NO_x conversion is still 87% in a 14-h night. If, on the other hand, N_2O_5 hydrolysis is slow, indirect losses via oxidative chemistry Reactions (R19) and (R20) dominate. Predicted summer and winter overnight NO_x loss would be 76% and 64%, respectively, assuming that all NO_3 reactions led to products other than NO_2 . In the section on “Field Determinations of the N_2O_5 Uptake Rate,” these rates of overnight NO_x loss will be estimated from a review of ambient measurements.

As N_2O_5 and NO_3 are always present in the same air mass, both direct and indirect losses have to be considered for a full picture of N_2O_5 chemistry. Thus, many of the instruments described in the third section, as well as the observations exemplified in the fourth section, measure and determine both tropospheric N_2O_5 and NO_3 levels at the same time.

TABLE 1
Summary of ambient N₂O₅ measurement techniques

Technique	Reference	Sensitivity	Averaging time (s)	Field deployment
Equilibrium calculation from measured NO ₂ and NO ₃				
DOAS	Atkinson et al. (1986) Platt and Stutz (2008)	NO ₃ : 2 pptv ^a	5 ^c	Phoenix, Arizona ^b TexAQS ^b
CE-DOAS	Meinen et al. (2010)	NO ₃ : 6.3 pptv ^a	0.001 ^c	N/A
Thermal conversion to NO ₃				
Pulsed CRDS	Brown et al. (2002) Dubé et al. (2006)	N ₂ O ₅ : 1 pptv	1	Boulder, Colorado NEAQS-ITCT 2004, TexAQS II
	Nakayama et al. (2008)	N ₂ O ₅ : 2.2 pptv	100	Toyokawa, Japan
Cw CRDS	Ayers and Simpson (2006)	N ₂ O ₅ : 2 pptv	25	Fairbanks, Alaska
OA CRDS	Ayers et al. (2005) Crowley et al. (2010)	NO ₃ : 1.4 pptv ^a N ₂ O ₅ : 4 pptv	4.6 3	Fairbanks, Alaska Kleiner Feldberg, Germany
Broadband CRDS	Ball et al. (2001) Bitter et al. (2005)	NO ₃ : 1 pptv ^a	100	NAMBLEX
CEAS	Langridge et al. (2008) Varma et al. (2009)	NO ₃ : 0.25 pptv ^a NO ₃ : 2 pptv ^a	10 5	RHaMBLe N/A
LIF	Wood et al. (2003) Matsumoto et al. (2005a)	N ₂ O ₅ : 6 pptv	600	Leuschner, California Tokyo, Japan
TD-CIMS	Slusher et al. (2004)	N ₂ O ₅ : 12 pptv	1	Boulder, Colorado
Direct ionization and detection of N ₂ O ₅				
ID-CIMS	Zheng et al. (2008)	N ₂ O ₅ : 30 pptv	10	MCMA-2006/MILAGRO
UW-CIMS	Kercher et al. (2009)	N ₂ O ₅ : 11 pptv	1	ICEALOT 2008

^aSee "Differential Optical Absorption Spectroscopy (DOAS)" section for conversion of NO₃ sensitivity to N₂O₅ sensitivity.

^bVertical profile field deployment measurements.

^cPath lengths (km) are shown instead of the averaging times here.

3. TECHNIQUES FOR MEASURING ATMOSPHERIC N₂O₅

There are three schemes for quantifying ambient N₂O₅. The first is the indirect calculation of N₂O₅ levels on the basis of its equilibrium with two directly measured species, NO₃ and NO₂. This method is common for open-path optical methods. In a closed-cell system, N₂O₅ can be measured by thermal conversion to NO₃ using a heated inlet and subsequent detection of the NO₃ molecules. This is common to in situ optical methods that are highly sensitive to NO₃, as well as to some mass-spectrometric instruments. These methods invariably measure the sum of NO₃ and N₂O₅, such that differentiation between the two must be accomplished either by a second measurement channel that does not include a thermal converter (thereby providing a measure of NO₃ alone) or by collocation of an NO₂ measurement such that the individual concentrations can be calculated from equilibrium. Finally, N₂O₅ may itself be directly ionized and detected via mass spectrometry. By exploiting these three basic schemes, several analytical techniques have been developed to quantify ambient N₂O₅. The techniques are described in this section and summarized in Table 1.

Many of the optical techniques described in this section have been intercompared during a recent measurement campaign at the SAPHIR environmental simulation chamber in Jülich, Germany (Fry et al. 2008; Rollins et al. 2009; Fuchs et al. 2010). Further description of those instruments and their accuracy will be given in a forthcoming publication (Apodaca et al. 2011). The chemical ionization techniques, however, did not participate in that campaign; consequently, Section 3.5 includes ambient intercomparisons of two chemical ionization instruments with a cavity ring-down spectrometer. Although a few other instrument intercomparisons have been conducted in ambient settings, a full description of them is beyond the scope of this review.

3.1. Differential Optical Absorption Spectroscopy (DOAS)

The first determinations of ambient N₂O₅ were made using long-path DOAS measurements of NO₂ and NO₃ in combination with Equation (1) (Platt et al. 1980; Atkinson et al. 1986). DOAS is a widely recognized detection technique for NO₃ in both remote and highly polluted environments. The advantage of DOAS lies in its ability to measure gaseous species such as

NO_2 and NO_3 without artifacts associated with wall losses and inlets. Since DOAS does not require calibration, it is considered an absolute analytical technique (Wang et al. 2006). During winter season though, the utility of DOAS for estimating N_2O_5 becomes quite limited by the low NO_3 levels. DOAS has been used to obtain long-term measurements of ground-level NO_3 (and NO_2) at several locations in Europe (e.g., Heintz et al. 1996; Allan et al. 1999; Geyer et al. 2001; Vrekoussis et al. 2007), from which N_2O_5 may be derived using the equilibrium relationship [Equation (1)]. More recently, long-path DOAS instruments have been used for vertical profiling in the lowest 300 m of the atmosphere, a height interval that is difficult to access otherwise (Section 4.2). Modern long-path DOAS instruments that measure tropospheric NO_2 and NO_3 are composed of a light source, such as a xenon arc lamp (Stutz et al. 2004), and a coaxial sending/receiving telescope that transmits a collimated beam of light to an array of quartz corner-cube reflectors positioned 2–6 km away from the instrument. The reflectors send the light back to the telescope where the narrow-band absorptions of various trace gases along the light path are analyzed with a spectrograph-detector system. For NO_3 , the window of wavelength is between 610 and 680 nm. To study the vertical distribution of trace gases in the lowest 100–300 m, three–five retroreflector arrays may be mounted at different altitudes. This setup allows the retrieval of concentration profiles with 2–40 m vertical resolution that are averaged across 2–6 km in the horizontal direction. This technique has been applied successfully in Houston and Phoenix (Stutz et al. 2004; Wang et al. 2006). Detection limits for NO_2 and NO_3 near the ground are typically 160 and 2 ppt, respectively, on a 5-km long light path (Platt and Stutz 2008). Applying Equation (1) at 1 ppbv of NO_2 , these values translate to N_2O_5 detection limits of 900 ppt at 255 K and 0.4 ppt at 310 K. The temporal resolution is between 10 and 20 min for each complete set of measurements.

3.2. Cavity Ring-Down Spectroscopy (CRDS)

A major breakthrough in the ambient measurement of N_2O_5 occurred over the past decade with advancements in CRDS. First reported by O’Keefe and Deacon (1988), the potential of CRDS for atmospheric trace gas measurements was recognized initially by successful measurements of ambient NO_2 in laboratory air (O’Keefe and Lee 1989). For further details on the technique, the reader is referred to numerous review articles (Busch and Busch 1999; Atkinson 2003; Brown 2003). King et al. (2000) demonstrated the first laboratory measurements of NO_3 by CRDS, and Brown et al. (2001, 2002) demonstrated the first in situ measurements of both NO_3 and N_2O_5 in ambient air by using CRDS. The latter system employs two mirrors of high finesse (mirror reflectivity, $R \geq 99.999\%$ at the strong, 662-nm NO_3 absorption band) and a tunable, pulsed dye laser system that injects the laser beam into the cavity directly through one of the end mirrors. This is referred to herein as pulsed CRDS. The concentration of the species of interest is determined by the decay of the laser intensity within the cavity, since the inten-

sity decay is directly proportional to the absorber concentration. Zero determination is obtained by periodic addition of NO to the inlet to destroy NO_3 via Reaction (R18). The combination of 662-nm optical extinction and chemical titration with NO has been shown to be an extremely specific method for NO_3 detection (Dubé et al. 2006). Thermal conversion of N_2O_5 to NO_3 in a second, heated channel provides simultaneous measurements of the sum of NO_3 and N_2O_5 . The measurement of N_2O_5 is obtained via the difference between the two channels. High sensitivity to NO_3 by direct absorption measurements using this single-wavelength method requires the removal of aerosol particles from the air sample via filtration, introducing potential inlet artifacts through loss of NO_3 and N_2O_5 on the filter. Artifacts due to the accumulation of reactive aerosol on the filter surface can be minimized by changing filters at regular intervals with an automated device (Dubé et al. 2006). Additional wall losses can be quantified and calibrated through standard additions and conversion of NO_3 or N_2O_5 to NO_2 , which is measured in a separate CRDS channel in the same instrument (Fuchs et al. 2008). Sensitivity and accuracy of the CRDS system has improved over the past decade. Currently, the system sensitivity is as good as 0.5 pptv for NO_3 and 1 pptv for N_2O_5 at a 1-Hz sampling frequency, with accuracy better than 12% (Dubé et al. 2006). Figure 4 shows a diagram of the instrument.

Pulsed CRDS has been deployed during several field campaigns on ships, aircraft, tall towers, and at ground sites. The instrument was first employed in March–April 2001 at a ground site in Boulder, Colorado (Brown et al. 2003b). The New England Air Quality Study (NEAQS) used the same setup in a ship-based application to measure N_2O_5 mixing ratios in a polluted marine environment off the United States east coast during the summer of 2002. In the 2004 International Consortium for Atmospheric Research on Transport and Transformation (ICARTT) study, spatial distributions and vertical profiles of N_2O_5 were obtained by installing a pulsed-CRDS system aboard an aircraft (Brown et al. 2006b, 2007a). In October 2004, Brown et al. (2007b) deployed the pulsed-CRDS system on a movable carriage on a 300-m tower in Boulder and reported vertical distributions of NO_3 and N_2O_5 with <1 m resolution. Nakayama et al. (2008) recently employed pulsed CRDS to obtain nighttime measurements of N_2O_5 in Toyokawa, Japan. The detection limit of that system was estimated to be 1.5 pptv for NO_3 and 2.2 pptv for N_2O_5 in a 100-s averaging time.

Simpson and coworkers (Simpson 2003; Ayers et al. 2005; Apodaca et al. 2008) have undertaken a parallel development of CRDS, using continuous wave (cw) diode lasers rather than a pulsed laser source. These cw CRDS systems have the advantage of much smaller size, weight, and power consumption, making them extremely versatile for use in field campaigns, as illustrated in Figure 5. However, the single CRDS channel restricts these systems to measuring only NO_3 or $\text{NO}_3 + \text{N}_2\text{O}_5$ depending on whether the thermal converter is bypassed or used. Thus, quantification of ambient N_2O_5 with this instrument requires a collocated measurement of NO_3 for a difference calculation or

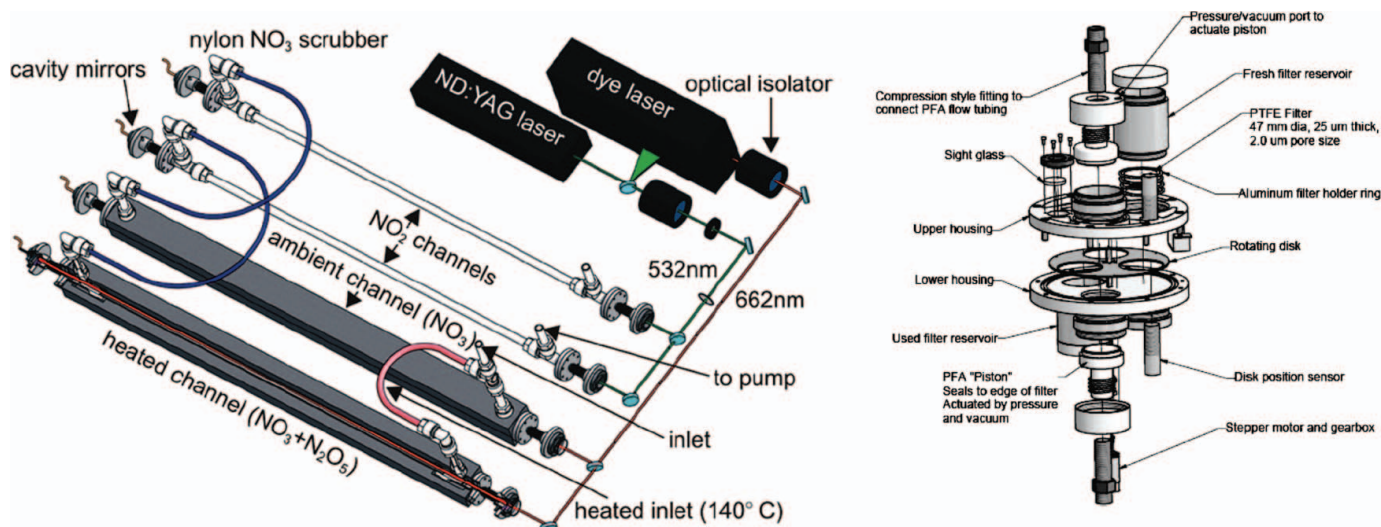


FIG. 4. Schematic of pulsed laser CRDS instrument for simultaneous measurement of NO_3 and N_2O_5 , with calibration based on conversion of these compounds to NO_2 , also measured by CRDS (Fuchs et al. 2008). Shown at right is the automated filter changer to prevent the accumulation of reactive aerosol (Dubé et al. 2006). Copyright 2008, American Chemical Society, and 2006, American Institute of Physics.

NO_2 and temperature for an equilibrium calculation. The first field-deployed cw CRDS near Fairbanks, Alaska, had a 2σ detection limit of 1.6 pptv for NO_3 or the sum of NO_3 and N_2O_5 in a 25-s averaging time (Simpson 2003). Ayers et al. (2005) reduced the detection limit to 1.4 pptv in 4.6 s by implementing an off-axis alignment that couples the cw diode laser light with a two-mirror system. Schuster et al. (2009) have also recently demonstrated detection of NO_3 and/or the sum of $\text{NO}_3 + \text{N}_2\text{O}_5$ in a single-channel system with a cw diode laser in an off-axis

configuration. This instrument achieves a sensitivity of 2 pptv in 5 s, with performance demonstrated during measurements at an environmental simulation chamber. Crowley and coworkers (2010) modified the system of Schuster et al. (2009) by introducing a second measurement cavity that allows for simultaneous measurement of NO_3 and N_2O_5 , similar to the dual-channel instruments described above. This instrument was deployed in southwestern Germany with an N_2O_5 detection limit of 4–5 pptv in a 3-s averaging time (Crowley et al. 2010).

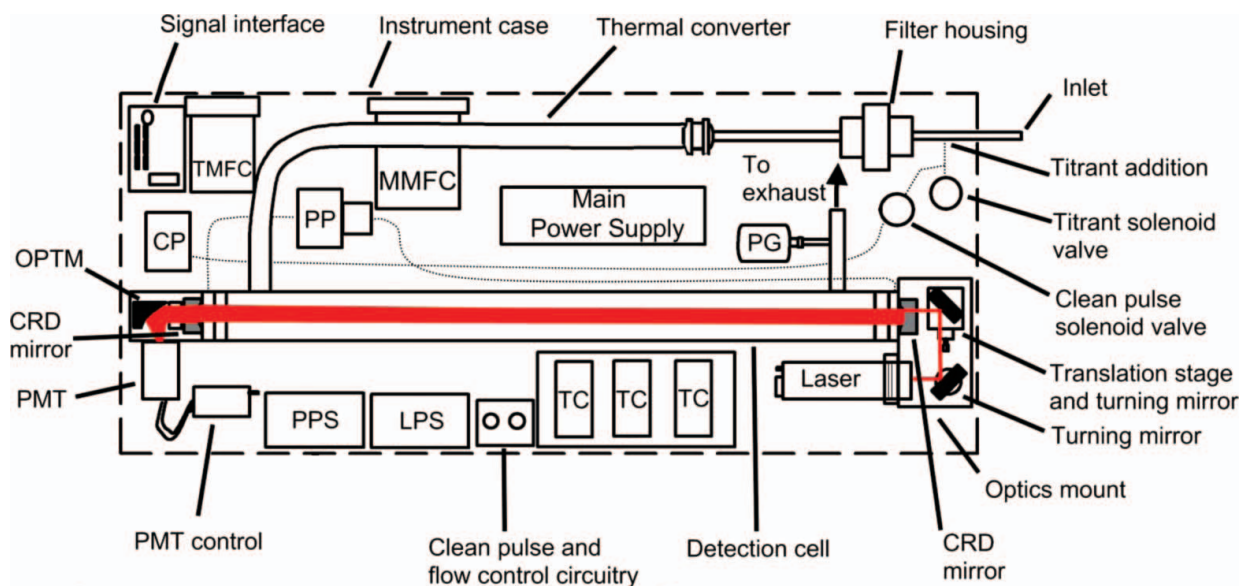


FIG. 5. Schematic of cw-CRDS instrument of Simpson (2003) for in situ detection of N_2O_5 . The instrument is designed for portability and includes a single CRDS channel with a thermal converter for measurement of the sum of NO_3 and N_2O_5 . TMFC: titrant mass flow controller; MMFC: main mass flow controller; PP: purge pump; CP: clean pulse pump; PG: pressure gauge; TC: temperature controller; LPS: laser power supply; PMT: photomultiplier tube; PPS: PMT power supply; OPTM: off-axis paraboloid turning mirror. Figure courtesy of William R. Simpson and Randy Apodaca.

3.3. Broadband Cavity Ring-Down and Cavity-Enhanced Absorption Spectroscopy (CEAS)

Concurrent with the development of single-wavelength CRDS instruments that measured NO_3 by optical extinction at the center of its strong 662-nm bands was the development of broadband optical absorption methods using high-finesse optical cavities that allow spectroscopic identification of NO_3 across a much wider wavelength range (Ball and Jones 2003). Because such instruments record a complete optical spectrum over the range where NO_3 absorbs strongly, they can be used to detect NO_3 without the need for zero determination via NO addition. Furthermore, they have commonly been operated without aerosol filtering, such that analytical uncertainties due to inlet losses are significantly reduced and those due to filter aging are eliminated. The initial demonstration of these instruments employed a pulsed dye laser system operating in a broadband mode (several tens of nm bandwidth), a grating spectrometer to disperse light at the output of an optical cavity, and a clocked charge-coupled device camera to record cavity ring-down transients within multiple wavelength bins (Ball et al. 2001; Bitter et al. 2005). This instrument demonstrated sensitivity to NO_3

of 1 pptv in a 100-s integration time on a single channel and was used for ambient measurements in comparison to long-path DOAS during the 2002 NAMBLEX campaign in Great Britain (Sommariva et al. 2007).

More recent developments of this instrumentation have employed light-emitting diodes (LEDs) in place of pulsed laser systems (Ball et al. 2004; Langridge et al. 2006, 2008) in a configuration known as CEAS, in which intensity, rather than a cavity time constant, is measured (Fiedler et al. 2003). These instruments require some additional calibrations of cavity mirror losses and are thus not absolute in the same sense as CRDS. Current versions of this instrument have achieved sensitivity to NO_3 or the sum of NO_3 and N_2O_5 in a single channel of 0.25 pptv (1σ) in a 10-s average, quite comparable to the single-wavelength CRDS instruments. The extreme simplicity and low cost of the LED makes these instruments attractive for field measurements. A schematic of the Langridge et al. (2008) instrument is shown in Figure 6. These instruments have been deployed in recent ground-based field intensives in Europe, such as the RHAMBLe (Reactive Halogens in the Marine Boundary Layer) in Roscoff, France, during 2006. The single-channel broadband instruments

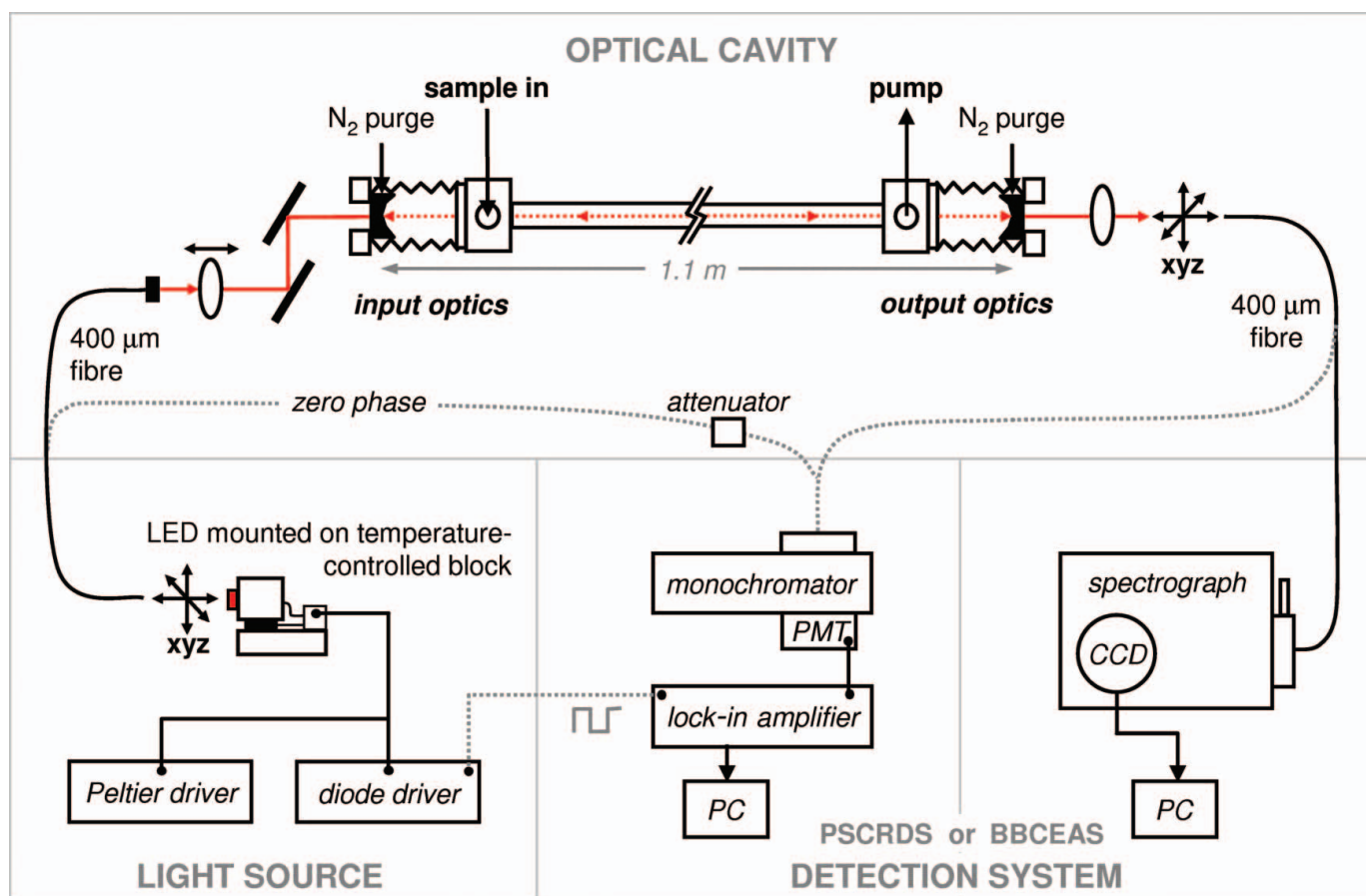


FIG. 6. LED-based cavity enhanced spectrometer for NO_3 or the sum $\text{NO}_3 + \text{N}_2\text{O}_5$ from Langridge et al. (2008). LED: light emitting diode; PMT: photomultiplier tube; PC: personal computer; CCD: charge-coupled device; PSCRDS: phase-shift cavity ring-down spectroscopy; BBCEAS: broadband cavity enhanced spectroscopy. Copyright 2008, American Institute of Physics.

described here have been designed to measure the sum of NO_3 and N_2O_5 and thus would currently require a collocated measurement of NO_2 and ambient temperature to determine NO_3 or N_2O_5 individually. However, dual-channel versions of these instruments that simultaneously measure NO_3 and the sum of NO_3 and N_2O_5 are straightforward and will likely be developed in the near future.

There are a number of additional CEAS instruments whose performance has been demonstrated in environmental simulation chambers and that are likely to be used in field measurements in the near future. These include arc-lamp-based instrumentation of Venables et al. (2006) and Varma et al. (2009), which have sensitivities to NO_3 of 2–4 pptv on 5–20 m open paths within an enclosed chamber with sampling times of 5 s to 1 min. Similarly, Meinen et al. (2010) reports NO_3 detection with an LED-based CEAS instrument using DOAS retrieval techniques (Platt et al. 2009), also referred to as CE-DOAS (Table 1), to achieve a sensitivity of 6.3 pptv to NO_3 on a 1-m base path. Because these instruments have been operated in open-path configurations, they are limited to measurements of NO_3 only and require collocated measurements of NO_2 and ambient temperature in order to determine N_2O_5 .

3.4. Laser-Induced Fluorescence (LIF)

LIF has been used for its accuracy and sensitivity in measuring a variety of atmospheric species, such as NO_2 and related nitrogen compounds (Thornton et al. 2000) and the HO_x radical family (Holland et al. 1995, Heard and Pilling 2003). The technique uses a laser to excite the species of interest to a higher electronic state, from which it may fluoresce with an efficiency determined by the photophysical properties of the molecule of interest. In general, the technique is extremely sensitive since photons emitted perpendicular to the excitation axis defined by the laser source can be collected and measured with high efficiency. It can also be a highly selective method since the fluorescent photons must satisfy two resonant conditions—the excitation from the laser itself and any optical filtering used to define the wavelength window over which the fluorescence is collected. The utility of LIF for particular trace gas measurements then depends on the accessibility and intensity of electronic transitions in regions where laser sources are available and the fluorescence yield.

The nitrate radical, with its strong and highly accessible visible absorption bands, would seem an ideal candidate for LIF detection of either this compound or N_2O_5 via thermal conversion, as described above. Wood et al. (2003) pioneered the use of this technique, demonstrating detection of NO_3 (or N_2O_5) in a prototype instrument that used a simple, inexpensive cw laser source (36-mW total power) that could be coarsely tuned around the strong 662-nm NO_3 absorption band. The system employed a multipass Herriot cell (approximately 40 passes), photon counting, and detection within a 700–750 nm spectral window to achieve optimal sensitivity and specificity to NO_3 . The most important contributions to the background signal (i.e.,

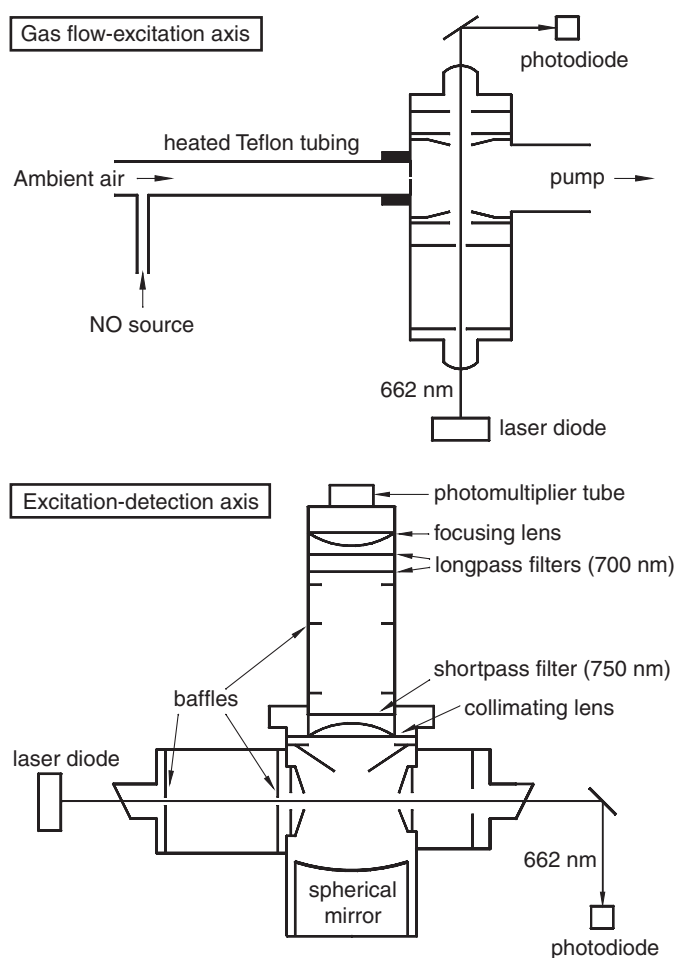


FIG. 7. LIF instrument for N_2O_5 measurement. Figure courtesy of Ezra Wood.

interferences) were due to scattering processes, such as Mie and Rayleigh scattering, scatter from the excitation laser itself, and fluorescence from NO_2 . The instrument sensitivity to NO_3 fluorescence was calibrated by conversion of NO_3 to NO_2 by reaction with NO [i.e., Reaction (R18)], followed by LIF measurement of NO_2 using a separate laser (with independent calibration of the NO_2 response). Figure 7 shows a schematic view of this instrument. A key finding was that the overall LIF sensitivity to NO_3 was some 1,600 times smaller than estimated based on comparison to similar instruments in use for NO_2 . These results implied that although the instrument was capable of selective detection of NO_3 or N_2O_5 , it likely had sensitivity sufficient for detection only of the latter, with a detection limit of 76 pptv (2σ) in a 60-s average.

Field measurements of N_2O_5 were demonstrated with this instrument during a wintertime campaign at a site on the East side of the San Francisco Bay area (Wood et al. 2005). The instrument was operated under conditions where the sum of NO_3 and N_2O_5 would be dominated by N_2O_5 , such that no NO_3 measurements were reported, and the sum was taken to be representative of N_2O_5 alone. To achieve higher sensitivity,

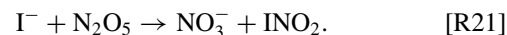
data were averaged to 1 h, with a detection limit of 6 pptv (2σ) in this time window. The study demonstrated lifetimes of N_2O_5 that were very short, in the range 5–30 min, implying very rapid wintertime conversion of NO_x to soluble nitrate via N_2O_5 hydrolysis.

Matsumoto et al. (2005a) developed an instrument for N_2O_5 detection, also based on thermal conversion (TC)/LIF. Their instrument operated somewhat differently and achieved somewhat improved detection sensitivity, albeit at a cost of greater experimental complexity. The key difference was the use of a higher powered (400 mW), pulsed laser system (10-kHz repetition rate) that enabled time-gated detection of the fluorescent signal to discriminate against scattered light from the excitation laser. The laser system was a pulsed Nd:YVO4/dye laser, which was large enough that the laser head itself was housed inside a laboratory, with a fiber-optic connection to the optical head where actual detection was carried out. They employed a dual-wavelength excitation scheme, with a signal wavelength resonant with the NO_3 band at 622.96 nm, and an off-resonant measurement at 618.81 nm. This scheme removed any potential interference from NO_2 fluorescence (as well as light scattering from aerosol), since the two wavelengths were chosen to have the same NO_2 absorption cross section to within 1%. As with the cw instrument of Wood et al., the instrument was calibrated against NO_2 fluorescence subsequent to the conversion of NO_3 found in the added excess NO. This instrument used the same reaction for recording an instrument zero, similar to the CRDS instruments described above. The detection limit of N_2O_5 in the system by Matsumoto et al. (2005b) is 6 pptv (1σ) over an integration time of 10 min. This would correspond to roughly 38 pptv (2σ) in 1 min, or approximately a factor of two better than the cw instrument of Wood et al. The pulsed-laser instrument has been used at field campaigns to measure $\text{NO}_3 + \text{N}_2\text{O}_5$ in the urban atmosphere near Tokyo, Japan (Matsumoto et al. 2005a) and on an island downwind of Tokyo (Matsumoto et al. 2006).

3.5. Chemical Ionization Mass Spectrometry (CIMS)

CIMS is a versatile and well-developed measurement technique used for a variety of trace gases (Huey et al. 1995). It is based on a soft and selective ionization process resulting from a reaction between a reagent ion and the compound of interest. After the molecule-ion reaction, ions expand through an orifice into a vacuum chamber for analysis by a mass spectrometer. The choice of reagent ion can lead to specificity for detecting particular compounds, while the soft ionization tends to produce simple mass spectra with relatively little fragmentation. Fehsenfeld et al. (1975) and Davidson et al. (1978) demonstrated that the gas-phase ion-molecule reaction between N_2O_5 and halide anions yields NO_3^- , which can be readily detected. Hanson and Ravishankara (1991) and later Abbatt and coworkers (Hu and Abbatt 1997; Thornton et al. 2003) applied this technique to study the heterogeneous uptake kinetics of N_2O_5 and ClONO_2 in the laboratory by using I^- as a reagent ion be-

cause it ionizes N_2O_5 with at most only a weak interference from HNO_3 .



The I^- reagent also ionizes NO_3 to NO_3^- , so the technique cannot differentiate between NO_3 and N_2O_5 . Iodide-based CIMS methods were first applied in field studies by Huey and coworkers (Slusher et al. 2004). Figure 8 illustrates the basic schematic of the apparatus. I^- is generated from a small flow of CH_3I passed over a ^{210}Po source. The ion-molecule reactions take place in a flow tube maintained at 20 Torr, followed by a collisional dissociation chamber (CDC), which consists of an electric field that accelerates the product ions through a gas at <0.5 Torr to break up clusters and simplify the mass spectrum. Because the principal application of this instrument is the speciation of peroxy acetyl nitrate (PAN) compounds, which requires thermal dissociation (TD) of PAN prior to the ion-molecule reaction, the instrument uses a heated inlet (TD-CIMS). Calibration of the instrument for N_2O_5 is estimated from the known sensitivity to PAN. Although the TD step is not technically required to measure the sum of $\text{NO}_3 + \text{N}_2\text{O}_5$ (because I^- reacts with either species to produce NO_3^-), it is convenient since the addition of NO to the inlet provides a zero in the same fashion as the optically based N_2O_5 instruments described above. A principal drawback of this TD-CIMS method is that although I^- is a reasonably specific reagent ion for either NO_3 or N_2O_5 , there is significant and potentially variable background noise at the 62 mass (NO_3^- amu). In some regions of the atmosphere, there may be additional interference from the reaction of I^- with ClONO_2 or BrONO_2 , which also yields NO_3^- . The estimated TD-CIMS detection limit for $\text{NO}_3 + \text{N}_2\text{O}_5$ is 12 pptv in 1 s. The instrument was deployed for field measurements in Boulder, Colorado, during October 2002 (Slusher et al. 2004), which revealed high variability in N_2O_5 with an increasing trend throughout the night and rapid depletion after sunrise. A field comparison between the TD-CIMS and pulsed CRDS (Dubé et al. 2006), in which both instruments were calibrated independently, exhibited high correlation between the instruments (under conditions where the sum of $\text{NO}_3 + \text{N}_2\text{O}_5$ was dominated by N_2O_5), though the TD-CIMS consistently reported 30% less N_2O_5 than CRDS (see Figures 9a and 9b).

Ion-drift CIMS (or ID-CIMS) is a similar apparatus that has also been applied to detect N_2O_5 in the atmosphere by using I^- as a reagent ion. The basis of this technique is the use of an ion drift tube in which the time for ion-molecule reactions to occur is well constrained, such that the signal can be calibrated based on a known ion-molecule reaction rate coefficient (Fortner et al. 2004). Zheng et al. (2008) demonstrated this method for N_2O_5 (and HNO_3 , via separate reagent ion chemistry) detection during field measurements in Mexico City in 2006. The instrument was similar in many respects to that of Slusher et al. (2004), except that it did not employ a heated inlet and so it measured N_2O_5 directly via Reaction (R9). The instrument was

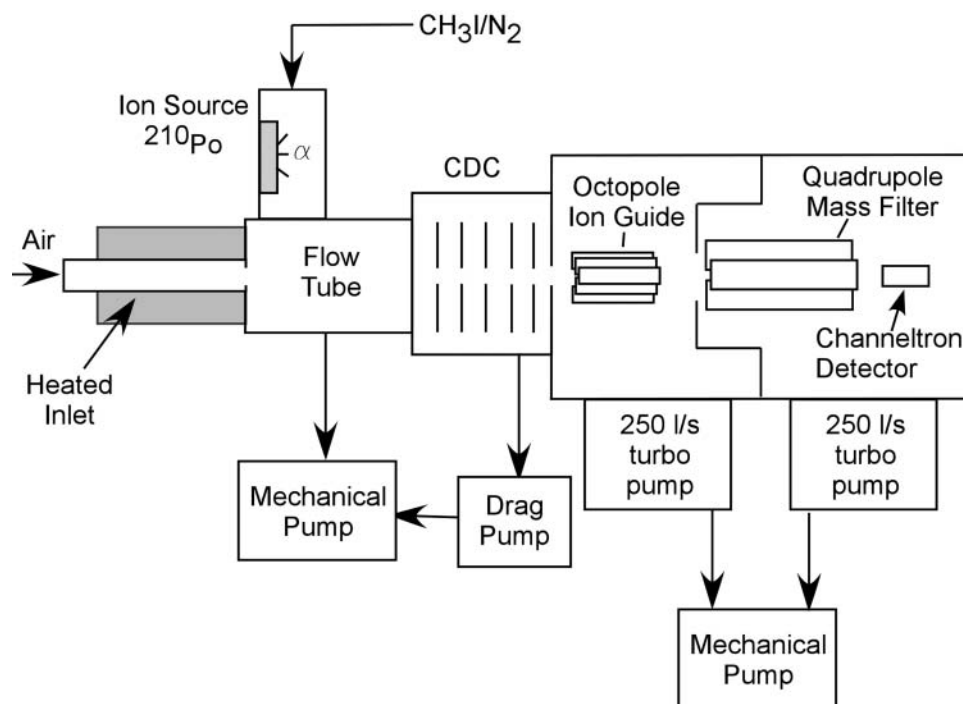
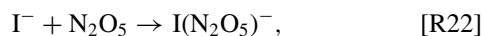


FIG. 8. Chemical ionization mass spectrometer for N₂O₅ detection. CDC: collisional dissociation chamber (Adapted from Slusher et al. 2004).

zeroed by shunting the ambient air flow through a heated tube to dissociate N₂O₅ into NO₃, followed by addition of NO to the inlet. Calibrations were performed separately in the laboratory from prepared samples of N₂O₅ that were characterized by a commercial ultraviolet spectrometer. The reported detection limit is 30 pptv (3σ) for a 10-s averaging time (Zheng et al. 2008).

The University of Washington CIMS (UW-CIMS) is similar to the TD-CIMS, but it has an ion–molecule reaction region (IMR) in place of the heated inlet and flow tube. By removing the heated inlet, N₂O₅ molecules remain intact and can be directly measured rather than the sum of NO₃ + N₂O₅. In the IMR, neutral molecules are allowed to react with I[−] ions for ~70 ms (Kercher et al. 2009). The novelty of this technique is its lack of high and variable chemical interferences because it detects iodide-containing clusters,



rather than nitrate anions. The I(N₂O₅)[−] cluster is an intermediate in the ion–molecule reaction leading to nitrate formation and must be detected before it dissociates so that the instrument can distinguish N₂O₅ from other sources of NO₃[−] [compare Reactions (R21) and (R22)]. Formation of this cluster is optimized in the UW-CIMS by tuning the flow-tube pressures and the CDC voltage. This instrument was deployed on a research vessel at the International Chemistry in the Arctic Lower Troposphere (ICEALOT) campaign and demonstrated a detection limit of ~5 pptv. Prior to the ICEALOT campaign,

the UW-CIMS performance was tested against pulsed CRDS in Boulder, Colorado. Results of that intercomparison, plotted in Figures 9c and d, show excellent agreement between the two measurement techniques. However, unlike the intercomparison plotted in Figures 9a and b, the UW-CIMS calibration was based partly on a direct comparison with the CRDS calibration source (Fuchs et al. 2008).

4. FIELD OBSERVATIONS OF N₂O₅

Determinations of ambient N₂O₅ levels have been available for at least 25 years by equilibrium calculations from simultaneous measurements of NO₃ and NO₂ by long-path DOAS (Atkinson et al. 1986), and direct in situ observations have been available since 2001 (Brown et al. 2001). Since a comprehensive review of all field measurements is beyond the scope of this article, we summarize several key findings in this section from recent field studies using in situ methods or vertically resolved DOAS. These include recent advances in our knowledge of N₂O₅ chemistry in cold climates, its vertical distribution in the lower atmosphere, and the rate of and reaction products from N₂O₅ heterogeneous hydrolysis.

4.1. N₂O₅ Levels in the Atmosphere

Since the discovery of the NO₃ radical in the atmosphere of Southern California in 1979, studies have speculated about the levels of N₂O₅ and its potential impact on the nocturnal NO_x budget. Calculated equilibrium N₂O₅ mixing ratios in these initial studies were reported at 1.5–14 ppb in polluted air (Platt et al. 1980; Atkinson et al. 1986), while long-term measurements

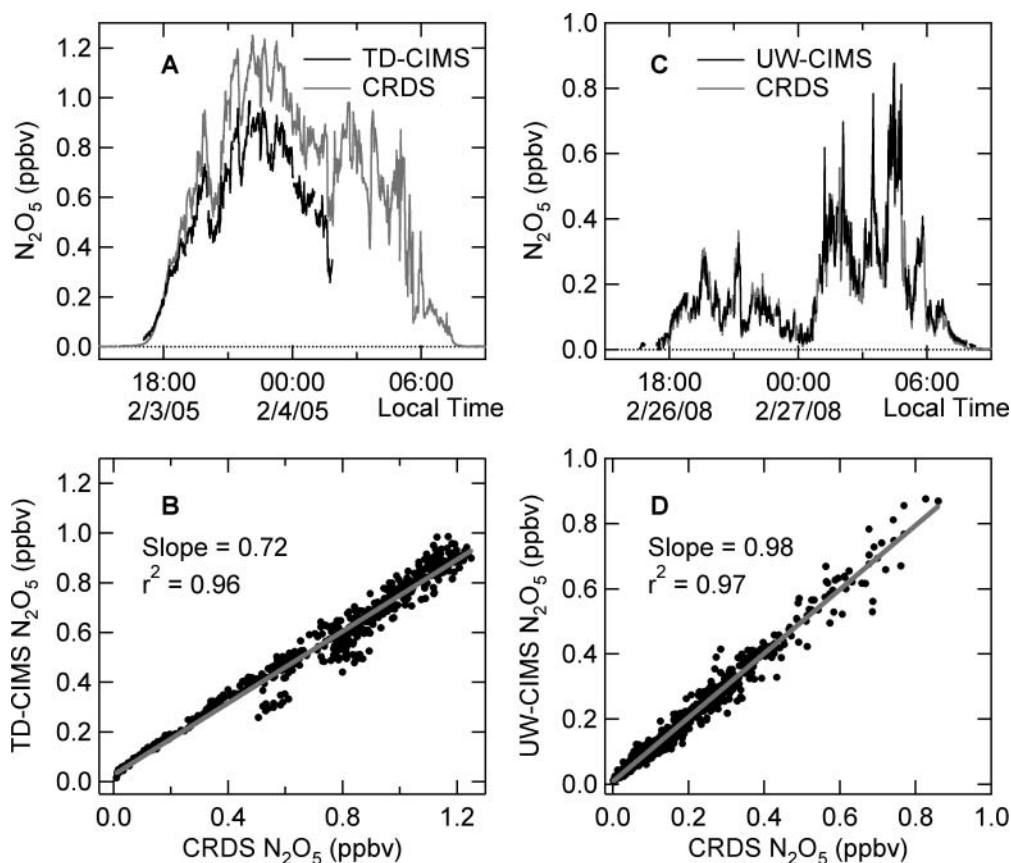


FIG. 9. Data from two separate intercomparisons between CIMS and CRDS instruments sampling ambient air in Boulder, Colorado. (a, b) A February 2005 comparison between the TD-CIMS instrument (Slusher et al. 2004), based on NO_3^- , with the pulsed CRDS instrument (Dubé et al. 2006). The TD-CIMS data have an arbitrary offset that has been subtracted. (c, d) A February 2008 comparison of the UW-CIMS, based on $\text{I}^-\text{N}_2\text{O}_5^-$, with the same pulsed CRDS instrument. Data in both comparisons are plotted at approximately 1-min time resolution. TD-CIMS data courtesy of Greg Huey, Frank Flocke, and Aaron Swanson. UW-CIMS data courtesy of James Kercher and Joel Thornton.

in marine and continental areas in Europe found up to 3.8 ppb of N_2O_5 (Heintz et al. 1996; Geyer et al. 2001). Analyses of long-term data show distinct seasonal variations in the N_2O_5 loss pathways, with direct loss of N_2O_5 most important in winter and indirect loss via NO_3 of greatest importance in summer (Heintz et al. 1996; Allan et al. 1999; Geyer et al. 2001; Vrekoussis et al. 2007). Other shorter field observations, including the in situ measurements by Brown et al. (2003b, 2006b, 2007b), have generally confirmed the presence of a few ppb of N_2O_5 in the troposphere. Atmospheric observations have confirmed that the equilibrium Reaction (R2) holds for most conditions during summertime (Perner et al. 1985; Brown et al. 2003b; Osthoff et al. 2007; Crowley et al. 2010), suggesting that the larger body of data that inferred N_2O_5 from NO_2 and NO_3 is likely valid in warm conditions. Field data have also given important insights into atmospheric N_2O_5 chemistry, including the role of N_2O_5 uptake as an NO_x loss mechanism (Geyer et al. 2001; Stutz et al. 2004) and the N_2O_5 lifetime during summer (see Section 4.3).

In cold environments (e.g., high latitudes, high altitudes, and/or winter months), it becomes difficult to infer N_2O_5 mixing

ratios from NO_3 because the equilibrium reaction (R2) shifts strongly in favor of N_2O_5 (Figure 1) and the NO_3 levels often drop below detection limits. Under such conditions, in situ measurement techniques are critical to determining the N_2O_5 mixing ratio. At a polluted Arctic site during winter, Apodaca et al. (2008) observed N_2O_5 levels as high as 50–100 ppt. This study and a previous study at similar conditions by Ayers and Simpson (2006) estimated N_2O_5 lifetimes from below 1 min up to 4 h. These short lifetimes are indicative of a fast, direct loss of N_2O_5 because NO_3 levels are so low at cold temperatures. Apodaca et al. (2008) argue that N_2O_5 is predominately lost on snow or ice particles. The dark reaction of NO_2 with O_3 to produce N_2O_5 is a more important mechanism for NO_x loss in wintertime, when daytime NO_x loss, driven by the reaction of NO_2 with photochemically produced OH radicals, is significantly reduced.

4.2. N_2O_5 Vertical Distribution

Weak vertical mixing in the nighttime boundary layer (NBL), which often is shallower than ~ 200 m, leads to an accumulation of freshly emitted pollutants (e.g., NO) near the surface.

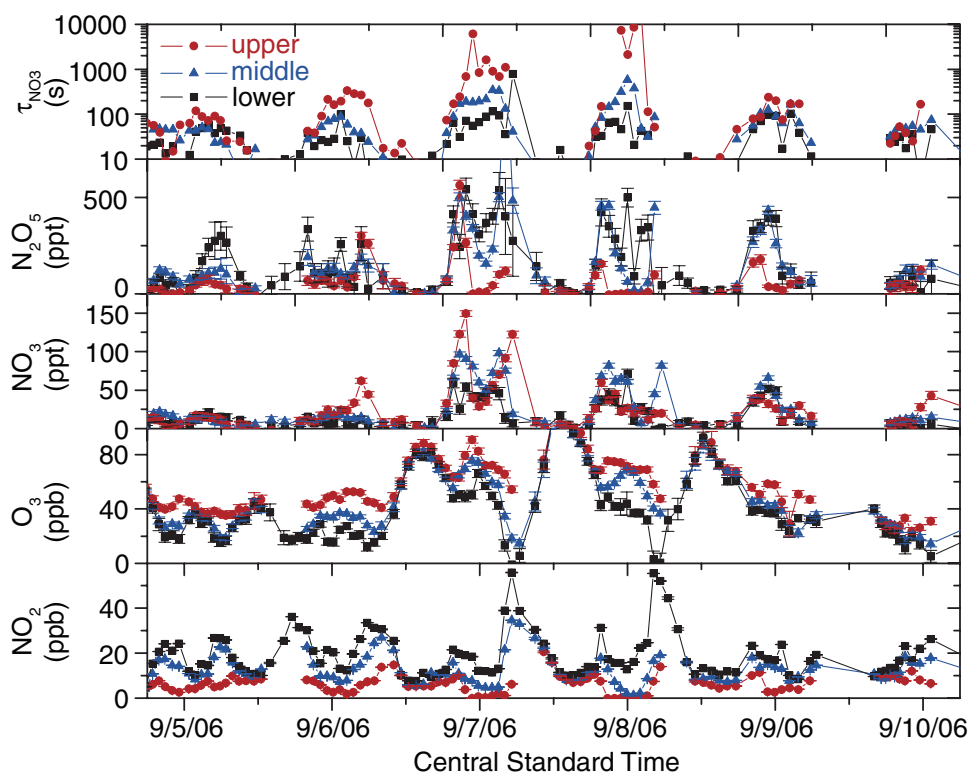


FIG. 10. Vertical profiles of various atmospheric trace gases in Houston, Texas, measured by long-path DOAS in September 2006. Data are shown in three height intervals: lower (20–70 m), middle (70–130 m), and upper (130–300 m). (Adapted from Stutz et al. 2009.)

This can result in efficient titration of NO₃ by NO [Reaction (R18)], as well as titration of O₃ by NO, leading to complex vertical concentration gradients in NO₃, N₂O₅, and their source gases, as well as a very altitude-dependent chemistry. By using DOAS measurements of scattered solar light during sunrise, von Friedeburg et al. (2002) were able to construct one of the first vertical profiles of tropospheric NO₃. It exhibited a maximum at ~300 m, the height where nocturnal jets are capable of carrying N₂O₅ overnight to downwind distance of ~300 km. Since their technique did not allow the simultaneous measurement of NO₂ profiles, equilibrium N₂O₅ concentration profiles could not be derived. Stutz et al. (2004) used long-path DOAS on several vertically arranged light paths to simultaneously measure NO₃ and NO₂ and calculated equilibrium N₂O₅ profiles in the lowest 120 m of a suburban atmosphere near Houston during the Texas Air Quality Study (TexAQS) in 2000. For several nights, NO₃ and N₂O₅ exhibited negligible levels at the surface but reached 100 and 300 ppt, respectively, between 100 and 120 m altitude. Other studies in the urban NBL reveal similar vertical profiles (e.g., Wang et al. 2006).

Figure 10 presents sample concentration profiles of NO₂ and NO₃ in three height intervals over a 1-week period during the second TexAQS in 2006 (Stutz et al. 2009). The measurements were taken in coastal regions where strong internal boundary layers form, characteristic of areas near large bodies of water.

During all nights, NO₂ exhibits higher levels near the surface and lower mixing ratios aloft, while ozone shows the opposite behavior. In general, NO₃ levels in the lower NBL (20–130 m altitude) show the expected profile of low NO₃ near the surface due to Reaction (R18), while N₂O₅ tends to be lowest above the NBL due to the very low NO₂ mixing ratios in the residual layer (see gray circles in the N₂O₅ panel of Figure 10). Peak levels of calculated steady-state N₂O₅ were observed mostly in the lower or middle height interval, reaching up to 500 pptv. Such drastic variations in N₂O₅ levels show a clear vertical gradient during many nights of the measurement campaign. The vertical profile of N₂O₅ is further complicated due to its dependence on aerosol concentration profiles [Equation (4)], RH [via Reaction (R3)], and the vertical temperature gradient [Equation (1)]. Vertical profiles of N₂O₅ within and above the NBL were also observed by in situ measurements at a tall tower in Boulder, Colorado (Brown et al. 2007b), and also showed strong vertical gradients. In this study, such gradients were particularly sharp within the surface layer (approximately the lowest 20 m), again most likely due to Reaction (R18). These observations showed large variability, with N₂O₅ maxima up to 2 ppbv and distinct layering of N₂O₅ at certain times during the night.

The recent development of sensitive in situ techniques to measure NO₃ and N₂O₅ allowed further investigation of N₂O₅ chemistry in the residual layer and free troposphere. Aircraft

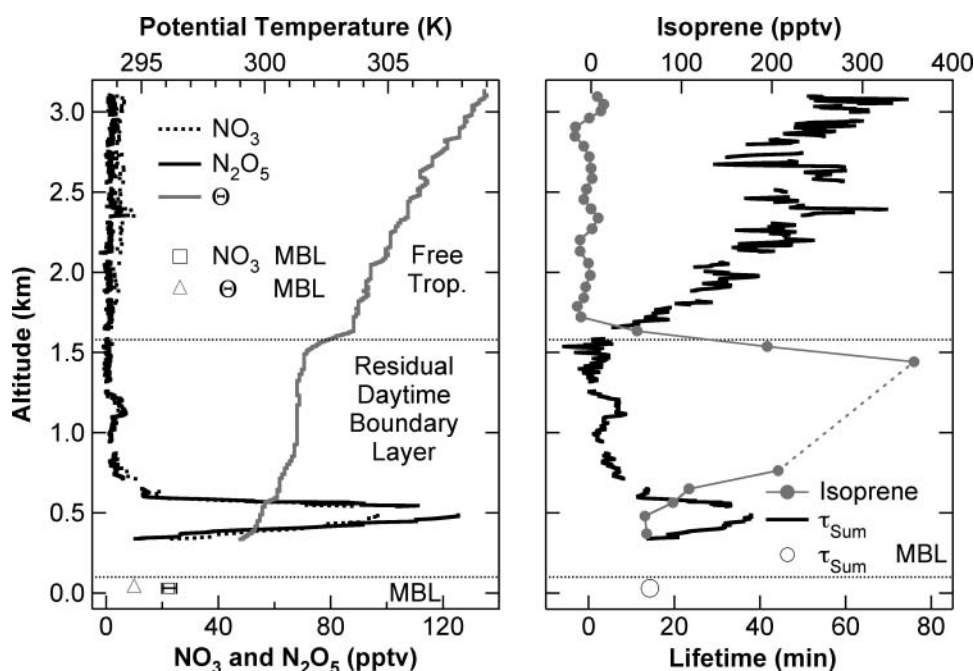


FIG. 11. Aircraft vertical profiles of NO_3 , N_2O_5 (measured by CRDS), potential temperature (Θ), isoprene, and $\text{NO}_3 + \text{N}_2\text{O}_5$ lifetime (τ_{Sum}), and surface-level measurements of NO_3 (measured by DOAS), Θ , and τ_{Sum} off the New Hampshire coastline in August 2004 (adapted from Brown et al. 2007a). Different symbols are used to distinguish measurements made in the marine boundary layer (MBL) from those collected in the aircraft, as indicated in the legends.

data, sometimes in combination with surface observations, often show an increase of NO_3 and N_2O_5 with altitude in the lowest 500 m of the nocturnal atmosphere, but the direction of the gradient may be reversed due to variations in pollutant compositions and the stability of the atmosphere. Figure 11 shows one example profile taken near the New Hampshire coastline in August 2004. A DOAS instrument with a light path between two islands provided surface-level measurements of NO_3 (and calculated N_2O_5 , not shown). The aircraft navigated a spiral ascent to just over 3 km. The altitude dependence of the $\text{NO}_3 + \text{N}_2\text{O}_5$ lifetime [$\tau_{\text{Sum}} = [\text{NO}_3 + \text{N}_2\text{O}_5] / (k_1 [\text{NO}_2][\text{O}_3])$], where k_1 is the reaction rate coefficient of Reaction (R1)] shows three distinct chemical regimes correlated with the layered structure in the potential temperature profile, including (1) the marine NBL, likely dominated by reactions of NO_3 with DMS emitted from the ocean (R19); (2) the residual daytime boundary layer, containing considerable level of isoprene (hence high rates of an indirect N_2O_5 loss pathway) advected offshore from the land surface; and (3) the free troposphere, where NO_3 and N_2O_5 have small production rates but long lifetimes. Even within the residual daytime boundary layer, additional stratification and concentration gradients are evident, such as the distinct plume near 500-m altitude (Brown et al. 2007a). On the basis of observed enhancements in tracers such as CO and SO_2 (not shown), this plume appears to have been of urban origin. It was likely not measured at surface level in this profile due to the interaction of the marine and terrestrial boundary layers in this region.

4.3. Field Determinations of the N_2O_5 Uptake Rate

Recent advances in the measurement of N_2O_5 from mobile platforms and direct measurements of N_2O_5 reactivity have provided new insight into the efficiency of its heterogeneous uptake in different parts of the troposphere and the factors governing that process. Brown et al. (2006a, b, 2009) analyzed in situ aircraft measurements of N_2O_5 by pulsed CRDS from the ICARTT campaign (Fehsenfeld et al. 2006) and the second TexAQS (Parrish et al. 2008) to quantitatively determine $\gamma_{\text{N}_2\text{O}_5}$. The analysis method relies on the assumption of a steady state in NO_3 and N_2O_5 with respect to their production and loss (Brown et al. 2003a). Scaling of the steady-state lifetimes [or the ratio of measured concentrations to production rates from Reaction (R1)] with NO_x within discrete plumes transected by a ship or aircraft provides a method to separate and quantify individual sinks for either NO_3 or N_2O_5 (Brown et al. 2009). The first key result from these studies is that $\gamma_{\text{N}_2\text{O}_5}$ values show considerable spatial variability (Brown et al. 2006b). For example, during a nighttime flight across the northeast US during ICARTT, lifetimes of N_2O_5 were found to be short in Region I, but considerably longer in Region III along the East Coast (Figure 12). All of the variability in N_2O_5 lifetimes is attributable to the rate of N_2O_5 uptake rather than to indirect loss via NO_3 chemistry. Plumes sampled in Region I exhibited $\gamma_{\text{N}_2\text{O}_5}$ values of 0.02, while those along the East Coast had $\gamma_{\text{N}_2\text{O}_5} \leq 0.002$. This tenfold difference significantly affected nocturnal NO_x lifetimes in the two regions, with 93% predicted overnight NO_x loss in Ohio but only 50% in New England.

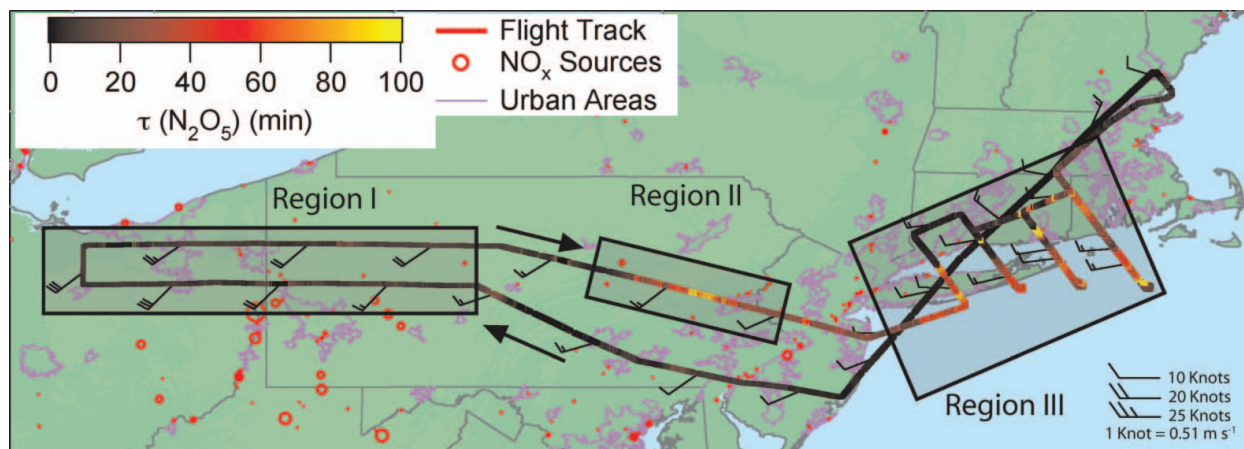


FIG. 12. Flight track of NOAA P-3 aircraft during a night in August 2004. The track is color coded by the observed lifetime of N₂O₅, $\tau(\text{N}_2\text{O}_5) = [\text{N}_2\text{O}_5]/k_1[\text{O}_3][\text{NO}_2]$, where k_1 is the rate coefficient of Reaction (R1). The map indicates both power-plant and urban NO_x sources and the areas over which the N₂O₅ uptake coefficient was observed to vary regionally (Brown et al. 2006b). Reprinted with permission from AAAS.

A second key result is that the values of $\gamma_{\text{N}_2\text{O}_5}$ seen in field studies were often smaller than those derived from laboratory measurements on pure inorganic salts, particularly along the East Coast, where neutral ammonium sulfate was the principal aerosol component. However, the field determinations are consistent with more recent laboratory measurements on mixed organic/inorganic systems discussed in Section 2.1. Determinations of $\gamma_{\text{N}_2\text{O}_5}$ from flights during the ICARTT campaign that had mixed organic/inorganic aerosol showed values generally smaller than 0.005.

The analysis of ICARTT and TexAQS aircraft measurements also indicated that the homogeneous hydrolysis rate coefficients recommended for Reaction (R3) and the corresponding third-order reaction (Section 2.1) on the basis of chamber studies appear to be too large (Brown et al. 2006b, 2009). While a recent field study in Germany supports this finding (Crowley et al. 2010), others have found that the inclusion of Reaction (R3) is necessary for models to accurately reproduce the observed NO₃ levels in a marine environment (Ambrose et al. 2007). Further field and laboratory work are needed to resolve these discrepancies.

More recent field measurements have investigated the reactivity of N₂O₅ directly using ambient flow-tube measurements in much the same manner that OH reactivity has been determined. These experiments were carried out using UW-CIMS detection of N₂O₅ (Section 3.5) coupled to a flow-tube reactor in which an artificial source of N₂O₅ is injected at the entrance, and the amount of N₂O₅ exiting the flow tube is measured in air samples that are alternately filtered and unfiltered of ambient particles (Bertram et al. 2009). Because the rate coefficient for N₂O₅ loss is relatively slow, the residence time in the flow tube must be much longer than needed for OH reactivity determinations in order to obtain a meaningful N₂O₅ loss rate coefficient. Although these measurements have so far been carried out

only at surface locations, their advantage over the steady-state method from aircraft transects is that they provide a larger number of measurements with which it is possible to statistically assess the factors that govern N₂O₅ reactivity. As an example, Figure 13 shows some values of $\gamma_{\text{N}_2\text{O}_5}$ determined via the UW-CIMS method during summer months at a low-RH (15%–45%) field site, Boulder, and a high-RH (61%–87%) field site, Seattle (Bertram et al. 2009). The comparison between measurements from the two sites not only showed strong dependence of $\gamma_{\text{N}_2\text{O}_5}$ on RH, the influence of varying aerosol organic-to-sulfate ratios is also prominent.

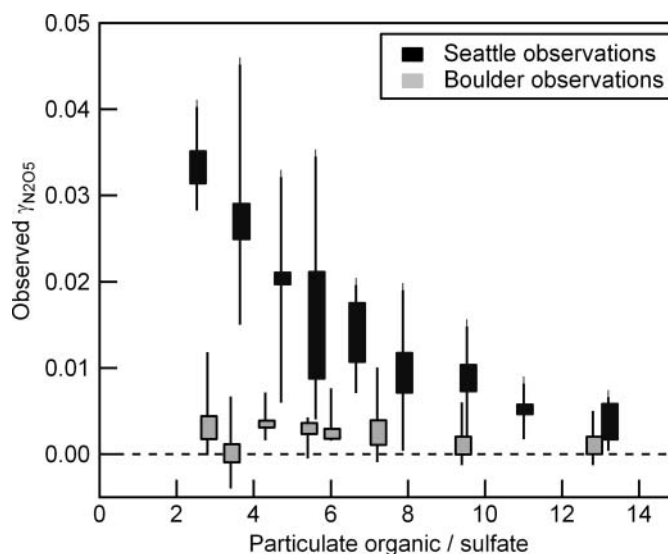


FIG. 13. Observed ambient $\gamma_{\text{N}_2\text{O}_5}$ with associated particulate organic-to-sulfate ratios at two field sites: Boulder (RH = 15%–45%) and Seattle (RH = 61%–87%). (Adapted from Bertram et al. 2009.)

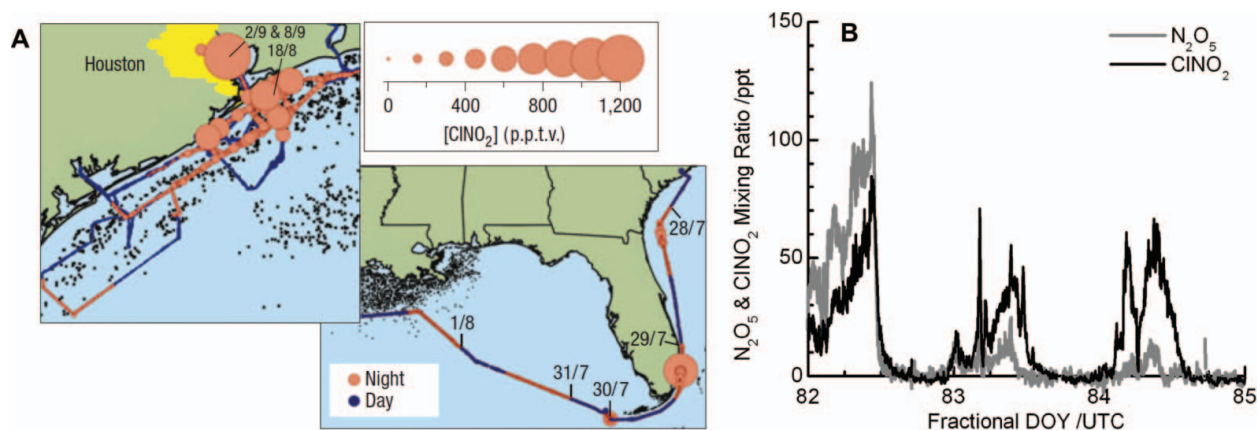


FIG. 14. (a) Map of the ship track of the R/V Brown during the second TexAQS campaign in 2006, color coded to show night and day, and size coded by the observed mixing ratio of ClNO_2 (Osthoﬀ et al. 2008). Reprinted by permission from Macmillan Publishers Ltd.: *Nature Geosciences*, copyright 2008. (b) The ClNO_2 and N_2O_5 mixing ratios measured from 22 March 2008 (DOY 82) through 24 March 2008 during the ICEALOT field campaign (From Kercher et al. 2009).

4.4. Field Determinations of ClNO_2 Production through N_2O_5 Heterogeneous Uptake

The product branching in N_2O_5 uptake is another key issue recently elucidated by field measurements. Most current atmospheric models that consider N_2O_5 heterogeneous uptake (Evans and Jacob 2005; Foley et al. 2010) treat HNO_3 as the only product [i.e., Reaction (R4)] and do not consider nitryl halide formation via Reaction (R12). Recent direct field measurements of nitryl chloride, ClNO_2 , have shown that it is produced in surprisingly large quantities from reactive uptake of N_2O_5 , implying that nighttime heterogeneous chemistry may be a larger source of tropospheric halogen activation than previously recognized (Osthoﬀ et al. 2008; Thornton et al. 2010). Previous estimates of the branching, based largely on the assumption that HNO_3 and ClNO_2 production occurred via uptake to submicron and supermicron (i.e., sea salt) aerosol, respectively, suggested that the production of ClNO_2 should be relatively modest, amounting to some tens of pptv in integrated overnight production in a polluted coastal environment (Aldener et al. 2006) or a global source of roughly 0.06 Tg Cl annually (Erickson et al. 1999). Recent ship-based measurements of N_2O_5 and, for the first time, ClNO_2 , during a 2006 campaign along regionally polluted areas of the US Gulf Coast, demonstrated that ClNO_2 occurs routinely at hundreds of pptv and occasionally ppbv levels (Figure 14a). Production of this order could be explained by invoking a relatively large yield of ClNO_2 on both sea-salt and submicron aerosols, suggesting that a significant fraction of total N_2O_5 consumption in polluted marine and coastal environments goes to the production of this gas-phase halogen compound. A revised global source estimate suggests annual Cl production of roughly 3.2 Tg in coastal and marine environments (Osthoﬀ et al. 2008) and up to 22 Tg if continental areas are also included (Thornton et al. 2010). Such estimates comprise a large fraction of current global budgets for tropospheric chlorine (Platt et al. 2004; Allan et al. 2007).

The 2006 results were from a late-summer campaign in a polluted subtropical marine boundary layer. More recent results from a second, ship-based campaign in late-winter 2008 (ICEALOT, <http://saga.pmel.noaa.gov/Field/icealot/index.html>) along coastal areas of the Northeast US and the North Atlantic have also shown efficient ClNO_2 production from shipping channels and coastal urban sources of NO_x . Similar simultaneous measurements of both N_2O_5 and ClNO_2 in the Long Island Sound and other locations downwind of the Northeast US corridor showed significant amounts of both compounds, with ClNO_2 levels generally exceeding those of N_2O_5 , as shown in Figure 14b (Kercher et al. 2009). Surprisingly, detectable levels of ambient ClNO_2 are not limited to coastal environments. Thornton et al. (2010) collected measurements near Boulder, Colorado, and found as much as 100–450 pptv of ClNO_2 in urban plume samples. These peaks correlated well with high mixing ratios of N_2O_5 occurring at night and early morning. Moreover, Thornton and coworkers estimated that more than half of the predicted ClNO_2 production in the United States occurs over land. Thus, ClNO_2 formation appears to be an important product of N_2O_5 heterogeneous uptake in a variety of environments, warranting further investigation of its production mechanisms.

5. MODELING TROPOSPHERIC N_2O_5

With increasing attention on laboratory and ambient measurements of N_2O_5 and its loss pathways, the modeling community has been working in parallel to understand the driving forces behind observed vertical gradients and regional distributions of N_2O_5 to quantify its impact on NO_x budgets and nitrate formation (Galmarini et al. 1997; Riemer et al. 2003). One of the major uncertainties in modeling ambient N_2O_5 is the parameterization of $\gamma_{\text{N}_2\text{O}_5}$. Results from both laboratory experiments and ambient measurements have shown that $\gamma_{\text{N}_2\text{O}_5}$ varies by several orders of magnitude, depending on the

prevailing meteorological conditions and aerosol composition (Sections 2.1 and 4.3). Also, the importance of homogeneous hydrolysis relative to heterogeneous hydrolysis is in question (Section 4.3). These uncertainties lead to further uncertainties in our understanding of the oxidizing capacity and pollutant levels in the troposphere, especially during nighttime. In particular, the NO_x and volatile organic compound (VOC) budgets as well as aerosol loading and composition are closely linked to N₂O₅ levels at night, as outlined in Figure 2 and Section 2.

5.1. Model Representation of N₂O₅ Uptake

When only homogeneous hydrolysis is considered, Emmerston and Evans (2009) showed that the reaction between N₂O₅ and H₂O contributes to a significant portion of uncertainties and discrepancies in inorganic chemistry among various gas-phase tropospheric chemistry schemes. Dimitroulopoulou and Marsh (1997) suggested adding a temperature dependence to the rate of Reaction (R3) based on early DOAS estimates of ambient N₂O₅ (Atkinson et al. 1986). This temperature-dependent function was implemented in the Particulate Matter Comprehensive Air Quality Model with eXtensions (PMCAMx) model for a study of seasonal pollutants variation in the eastern United States to avoid overprediction of HNO₃ level (Karydis et al. 2007). An early approach for modeling heterogeneous N₂O₅ hydrolysis in the troposphere was to treat it as a first-order loss reaction that produces HNO₃ (Russell et al. 1985; Chang et al. 1987; Hov et al. 1988), without explicitly modeling the aerosol surface area needed in Equations (3) and (4). Later, modeling studies incorporated N₂O₅ uptake by particle surfaces, assuming $\gamma_{\text{N}_2\text{O}_5}$ has a constant value of 0.1 (Dentener and Crutzen 1993; Pleim et al. 1995). Dentener and Crutzen (1993) found that by including the heterogeneous uptake of N₂O₅ and NO₃, the modeled yearly average global NO_x burden decreased by 50%. Similar results of up to 60% HNO₃ increase and 50% NO_x decrease were reported in a regional modeling study of the Houston-Galveston area where heterogeneous uptake with the same constant $\gamma_{\text{N}_2\text{O}_5}$ value of 0.1 was considered only on sulfate aerosols (Lei et al. 2004). However, Dentener and Crutzen (1993) forewarned that the laboratory measurements of $\gamma_{\text{N}_2\text{O}_5}$ varied by two orders of magnitude, and their estimate of 0.1 was an upper limit. In a more recent modeling study, Bauer and coworkers (2004) parameterized an RH-dependent $\gamma_{\text{N}_2\text{O}_5}$ function on mineral dust (0.02 at RH = 70% and 0.003 for RH = 30%) and showed that such uptake alone accounts for approximately 11% of global N₂O₅ loss, contributing as a major source of HNO₃.

There are several proposed mechanistic explanations for the variability of $\gamma_{\text{N}_2\text{O}_5}$ on the basis of aerosol composition and ambient conditions (refer to Section 2.1 for detailed discussions), but the associated rate constants for these reaction pathways remain highly uncertain. Thus, one way to develop $\gamma_{\text{N}_2\text{O}_5}$ as a function of environmental variables and aerosol mix is to rely on regression models to fit laboratory data (Evans and Jacob 2005; Davis et al. 2008). In an effort to improve model predictions

of the impact of N₂O₅ heterogeneous uptake while maintaining computational efficiency, Evans and Jacob (2005) implemented laboratory-based parameterizations of $\gamma_{\text{N}_2\text{O}_5}$ for five aerosol types: sulfate, organic carbon, black carbon, sea salt, and dust. For some of these aerosol types, $\gamma_{\text{N}_2\text{O}_5}$ was specified as a function of temperature and RH. Using the GEOS-CHEM global chemical transport model, they showed a 7% increase in the tropospheric NO_x burden when the updated parameterizations of $\gamma_{\text{N}_2\text{O}_5}$ were used in lieu of a constant value of 0.1. In addition, a significant increase in both global N₂O₅ and NO₃ concentrations was reported, resulting in an increase in ambient OH, which led to an overall increase in ozone concentration. More recently, Alexander et al. (2009) used the GEOS-CHEM model in conjunction with the N₂O₅ hydrolysis treatment presented by Evans and Jacob (2005) to model the oxygen isotopic composition ($\Delta^{17}\text{O}$) of nitrate and to estimate different pathways of nitrate formation. Their study showed that N₂O₅ hydrolysis contributes to 18% of the global annual-mean tropospheric inorganic nitrate and 74% at high northern latitudes over the continents and the Arctic.

Relative to the global-scale model simulations described above, more detailed treatments of N₂O₅ heterogeneous hydrolysis have been attempted in urban- and regional-scale modeling studies. In order to account for the suppression of $\gamma_{\text{N}_2\text{O}_5}$ when nitrate is present in the particles, Riemer et al. (2003) introduced the following parameterization:

$$\gamma_{\text{N}_2\text{O}_5, \text{mix}} = f \cdot \gamma_1 + (1 - f) \cdot \gamma_2, \quad [5]$$

with $\gamma_1 = 0.02$, $\gamma_2 = 0.002$, and

$$f = \frac{m_{\text{SO}_4}}{m_{\text{SO}_4} + m_{\text{NO}_3}}, \quad [6]$$

where m_{SO_4} and m_{NO_3} are the particle-phase mass concentrations of sulfate and nitrate, respectively. While Riemer et al. (2003) modeled the nitrate effect as a full order of magnitude suppression of $\gamma_{\text{N}_2\text{O}_5}$, studies on ammoniated salts showed that suppression to be between a factor of 1.3 and 5.2 (Davis et al. 2008).

Davis et al. (2008) developed a parameterization for $\gamma_{\text{N}_2\text{O}_5}$ on inorganic aerosols as a function of composition of the particle-phase state of the aerosol, RH, and temperature through a comprehensive examination of several laboratory studies (Table 2). The original $\gamma_{\text{N}_2\text{O}_5}$ parameterization on aqueous sulfate particles as a function of RH proposed by Davis et al. (2008) is similar to the parameterization by Evans and Jacob (2005). Due to the unconfirmed experimental values at high RH, however, Davis and coworkers recommended an “alternative” parameterization that limits the uptake probability above the threshold RH (46%) to $\gamma_{\text{N}_2\text{O}_5}(\text{RH} = 46\%)$. Davis and coworkers also recommended the introduction of temperature dependence for aqueous NH₄HSO₄ particles, where the uptake reaction probability can vary by an order of magnitude from 291 to 308 K,

TABLE 2

Summary of parameterization methods for N₂O₅ heterogeneous reaction probability ($\gamma_{\text{N}_2\text{O}_5}$) on particles of different compositions. Detailed description of the parameter constants for each method can be found in Sections 2.1 Direct Loss Pathways and 5.1 Model Representation of N₂O₅ Uptake

Particle Type	Parameterization	Parameters	Reference
Sulfate	$\gamma_{\text{N}_2\text{O}_5} = \alpha \times 10^\beta$	$\alpha = 2.79 \times 10^{-4} + 1.3 \times 10^{-4} \times \text{RH}$ $- 3.43 \times 10^{-6} \times \text{RH}^2$ $+ 7.52 \times 10^{-8} \times \text{RH}^3$ $\beta = 4 \times 10^{-2} \times (\text{T} - 294)$ (T < 282K) $\beta = -0.48(\text{T} < 282 \text{ K})$	Evans and Jacob (2005)
	$\gamma_{\text{N}_2\text{O}_5, x} = \frac{1}{1 + e^{-\lambda_x}}$, = AB, AS, d; $\gamma_{\text{N}_2\text{O}_5, AB}^* = \min(\gamma_{\text{N}_2\text{O}_5, AB}, 0.08585)$ $\gamma_{\text{N}_2\text{O}_5, AS}^* = \min(\gamma_{\text{N}_2\text{O}_5, AS}, 0.053)$ $\gamma_{\text{N}_2\text{O}_5, d}^* = \min(\gamma_{\text{N}_2\text{O}_5, d}, 0.0124)$ Ammonium bisulfate (AB): $\lambda_{AB} = \beta_{10} + \beta_{11}\text{RH}_{46} + \beta_{12}T_{291}$ Ammonium sulfate (AS): $\lambda_{AS} = (\beta_{10} + \beta_{20}) + \beta_{11}\text{RH}_{46}$ (crystallizes at RH < 32.8%) Dry ammoniated sulfates: $\lambda_d = \beta_{d0} + \beta_{d1}\text{RH} + \beta_{d2}T_{293}$	$\beta_{10} = -2.67270$ $\beta_{20} = -0.97579$ $\beta_{11} = 0.09553$ $\beta_{12} = -0.20427$ $\beta_{d0} = -6.13376$ $\beta_{d1} = 0.03592$ $\beta_{d2} = -0.19688$ $T_{291} = \max(\text{T} - 291, 0)$ $T_{293} = \max(\text{T} - 293, 0)$ $\text{RH}_{46} = \min(\text{RH} - 46, 0)$	Davis et al. (2008), Appendix
Nitrate	$\gamma_{\text{N}_2\text{O}_5, \text{AN}} = \frac{1}{1 + e^{-\lambda_{\text{AN}}}}$ $\gamma_{\text{N}_2\text{O}_5, \text{AN}}^* = \min(\gamma_{\text{N}_2\text{O}_5, \text{AN}}, 0.0154)$ $\lambda_{\text{AN}} = \beta_{30} + \beta_{31}\text{RH}$ $\gamma_{\text{N}_2\text{O}_5} = \left(\frac{1}{\alpha} + \frac{c_{\text{N}_2\text{O}_5} S}{4K_H RT k_V}\right)^{-1}$ $k_{\text{ref}} = k_5 \left(1 - \frac{k_a [\text{NO}_3^-]}{k_a [\text{NO}_3^-] + k_b [\text{H}_2\text{O}]}\right)$ $\gamma_{\text{N}_2\text{O}_5} = Ak' \left(1 - \frac{1}{\left(\frac{k_b [\text{H}_2\text{O}]}{k_a [\text{NO}_3^-}]\right) + 1 + \left(\frac{k_c [\text{Cl}^-]}{k_a [\text{NO}_3^-}]\right)}\right)$ $A = \frac{4V}{c_{\text{N}_2\text{O}_5} S} K_H$ $k' = \beta - \beta e^{-\delta [\text{H}_2\text{O}]}$	$\beta_{30} = -8.10774$ $\beta_{31} = 0.04902$ $V/S = 56 \text{ nm}$ for Fig. 15 $k_{\text{ref}} = 5 \times 10^6 \text{ s}^{-1}$ $k_a/k_b = 30$ $V/S = 37.5 \text{ nm}$ for Fig. 15 $\beta = 1.15 \times 10^6 \text{ s}^{-1}$ $\delta = 1.3 \times 10^{-1} \text{ M}^{-1}$ $k_b/k_a = 6.0 \times 10^{-2}$ $k_c/k_a = 29$	Davis et al. (2008) Griffith et al. (2009) Bertram and Thornton (2009)
Sulfate-nitrate	$\gamma_{\text{N}_2\text{O}_5, \text{mix}} = f \cdot \gamma_1 + (1 - f) \cdot \gamma_2$ $f = \frac{m_{\text{SO}_4}}{m_{\text{SO}_4} + m_{\text{NO}_3}}$ $\gamma_{\text{N}_2\text{O}_5, \text{aq}} = x_{AB} \gamma_{\text{N}_2\text{O}_5, AB}^* + x_{AS} \gamma_{\text{N}_2\text{O}_5, AS}^* + x_{AN} \gamma_{\text{N}_2\text{O}_5, AN}^*$ $\gamma_{\text{N}_2\text{O}_5, \text{dry}} = (x_{AB} + x_{AS}) \gamma_{\text{N}_2\text{O}_5, d}^* + x_{AN} \cdot \min(\gamma_{\text{N}_2\text{O}_5, d}^*, \gamma_{\text{N}_2\text{O}_5, AN}^*)$	$x_{AN} = 1 - (x_{AS} + x_{AB})$ $x_{AS} = \max(0, \min(1 - x_{AN}, \frac{[\text{NH}_4^+]}{[\text{NO}_3^-] + [\text{SO}_4^{2-}]} - 1))$ $x_{AN} = \frac{[\text{NO}_3^-]}{[\text{NO}_3^-] + [\text{SO}_4^{2-}]}$ m [mass]	Rierner et al. (2003) Davis et al. (2008)
Organic carbon	$\gamma_{\text{N}_2\text{O}_5} = \text{RH} \times 5.2 \times 10^{-4}$ (RH < 57%) $\gamma_{\text{N}_2\text{O}_5} = 0.03$ (RH \geq 57%)		Evans and Jacob (2005)
Organic coating	$\frac{1}{\gamma_{\text{N}_2\text{O}_5}} = \frac{1}{\gamma_{\text{N}_2\text{O}_5, \text{core}}} + \frac{1}{\gamma_{\text{N}_2\text{O}_5, \text{coat}}}$ $\gamma_{\text{N}_2\text{O}_5, \text{core}} = \gamma_{\text{N}_2\text{O}_5, \text{mix}}$ $\gamma_{\text{N}_2\text{O}_5, \text{coat}} = \frac{4RT H_{\text{org}} D_{\text{org}} R_c}{c_{\text{N}_2\text{O}_5} \ell R_p}$	$H_{\text{org}} D_{\text{org}} = 0.03 H_{\text{aq}} D_{\text{aq}}$ $H_{\text{aq}} = 5000 \text{ moles m}^{-1} \text{ atm}^{-1}$ $D_{\text{aq}} = 10^{-9} \text{ m}^2 \text{ s}^{-1}$	Anttila et al. (2006), Rierner et al. (2009)
Black carbon	$\gamma_{\text{N}_2\text{O}_5} = 0.005$		Evans and Jacob (2005)
Sea salt	$\gamma_{\text{N}_2\text{O}_5} = 0.005$ (RH < 62%) $\gamma_{\text{N}_2\text{O}_5} = 0.03$ (RH \geq 62%)		Evans and Jacob (2005)
Dust	$\gamma_{\text{N}_2\text{O}_5} = 0.01$		Evans and Jacob (2005)

as shown in Figure 15(top). The resulting $\gamma_{\text{N}_2\text{O}_5}$ function for sulfates at RH above 50% agrees well with the constant value of 0.02 suggested by Riemer et al. (2003) at 298 K. For dry ammoniated sulfate particles, a less RH-dependent $\gamma_{\text{N}_2\text{O}_5}$ function is recommended (Davis et al. 2008), and since the function does not surpass the upper limit of the experimental values even at high RH, no “alternative” parameterizations are necessary.

The parameterization of $\gamma_{\text{N}_2\text{O}_5}$ by Davis and coworkers on nitrate particles is also RH dependent, but no apparent temperature dependence was observed. It covers a range from 0.0005 (RH = 10%) to 0.0145 (RH = 80%). The constant value of 0.002 proposed by Riemer et al. (2003) only corresponds to the $\gamma_{\text{N}_2\text{O}_5}$ function by Davis et al. (2008) at 50% RH. Brown et al. (2009) compared these various parameterizations with field-based determinations of $\gamma_{\text{N}_2\text{O}_5}$. These field-based determinations yielded $\gamma_{\text{N}_2\text{O}_5}$ in the range of $0.5\text{--}6 \times 10^{-3}$, which is substantially smaller than current parameterizations used for atmospheric modeling (Riemer et al. 2003; Evans and Jacob 2005; Davis et al. 2008). Moreover, a dependence of $\gamma_{\text{N}_2\text{O}_5}$ on variables such as RH and aerosol composition was not apparent in the determinations.

The coating of particles by organic material also inhibits N₂O₅ uptake (Anttila et al. 2006) and has been suggested as a possible explanation for field observations of suppressed N₂O₅ uptake (Brown et al. 2006b). To quantify the effect of organic coatings, Anttila et al. (2006) treated the reactive uptake of gaseous N₂O₅ by aerosols using a resistor model. Their study considered coatings that are thick enough to be considered bulk absorbing phases, in contrast to monolayers or submonolayers that had been investigated previously. Anttila et al. (2006) presented the reaction probability in the form of

$$\gamma_{\text{N}_2\text{O}_5, \text{coat}} = \frac{4RT H_{\text{org}} D_{\text{org}} R_c}{c_{\text{N}_2\text{O}_5} \ell R_p}, \quad [7]$$

where R is the universal gas constant, T is the temperature, H_{org} is the Henry constant of the organic material, and D_{org} is the diffusion coefficient of the organic material; R_p , R_c , and ℓ are the radius of the particle, the radius of the core, and the thickness of the coating, respectively. As expected, the thicker the coating, the smaller the uptake of N₂O₅ by the particle. All the recommended values for constant variables are summarized in Table 2.

Combining both the nitrate- and organic-coating effects, Riemer et al. (2009) arrived at the following parameterization:

$$\frac{1}{\gamma_{\text{N}_2\text{O}_5}} = \frac{1}{\gamma_{\text{N}_2\text{O}_5, \text{core}}} + \frac{1}{\gamma_{\text{N}_2\text{O}_5, \text{coat}}}. \quad [8]$$

In a 3-D model, Riemer et al. (2009) estimated the maximum effect of organic coatings by assuming that all available secondary organic material contributed to the coating. That analysis showed that organics could suppress N₂O₅ uptake significantly,

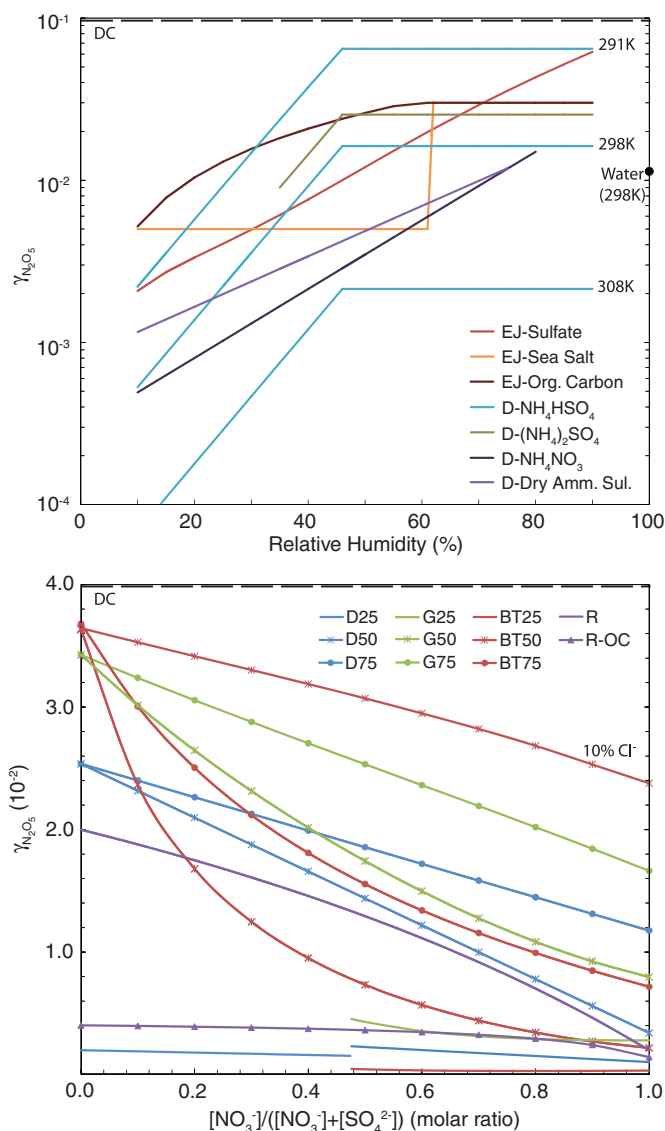


FIG. 15. N₂O₅ uptake reaction probability ($\gamma_{\text{N}_2\text{O}_5}$) parameterizations as a function of RH (top) and NO₃⁻ content (bottom) recommended by DC-Dentener and Crutzen (1993), EJ-Evans and Jacob (2005), D-Davis et al. (2008), G-Griffiths et al. (2009), BT-Bertram and Thornton (2009), R-Riemer et al. (2003), and R-OC-Riemer et al. (2009). Temperature of 298 K is used for all temperature-dependent parameterizations, unless otherwise specified. RH values for [NO₃⁻]-dependent $\gamma_{\text{N}_2\text{O}_5}$ parameterization (bottom) are indicated by the plot legend in percentage (e.g., D25 corresponds to 25% RH). The liquid water content of the aerosol, [H₂O], for Griffiths et al. (2009) and Bertram and Thornton (2009) is calculated with the aerosol inorganics model (AIM, <http://www.aim.env.uea.ac.uk/aim/aim.php>). The crystallization relative humidity (CRH) is determined according to the recommendation given by Martin et al. (2003), below which the gamma values are omitted for Griffiths et al. (2009) and Bertram and Thornton (2009) parameterizations, and dry particle equation is used for Davis et al. (2008). [Cl⁻] is not taken into consideration for $\gamma_{\text{N}_2\text{O}_5}$ parameterization by Bertram and Thornton, except for one case (BT50) where 10% of the total molar anion concentrations ([NO₃⁻] + [SO₄²⁻] + [Cl⁻]) is [Cl⁻] at 50% RH. For the organic-coating (OC) case (Riemer et al. 2009), particle size of 100 nm with 10% coating thickness is assumed. All parameterization constants used are summarized in Table 2.

reducing particle-phase nitrate concentrations by as much as 20%. They found the maximum effect of organics when nitrate concentrations are low and relatively little effect at high nitrate loadings. The parameterization by Riemer et al. (2009) leads to significantly lower reaction probabilities for N_2O_5 than the work presented by Davis et al. (2008).

A mechanistic parameterization of the nitrate effect in sulfate particles as a function of molar nitrate and water concentration was proposed by Griffiths et al. (2009),

$$k_r = k_{\text{ref}} \left(1 - \frac{k_a[\text{NO}_3^-]}{k_a[\text{NO}_3^-] + k_b[\text{H}_2\text{O}]} \right), \quad [9]$$

where $k_{\text{ref}} = 5 \times 10^6 \text{ s}^{-1}$ and the ratio k_a to k_b is 30. The effective $\gamma_{N_2O_5}$ for the proposed reaction constant, k_r , can be calculated by using the characteristic equation for reactive uptake coefficient on submicron aerosol when uptake is limited by the aerosol volume,

$$\gamma_{N_2O_5} = \left(\frac{1}{\alpha} + \frac{c_{N_2O_5} S}{4K_H R T k_r V} \right)^{-1}, \quad [10]$$

where α is the mass accommodation coefficient, K_H is the Henry's law constant for N_2O_5 in water, R is the universal gas constant, and V is the aerosol volume. Their results also indicated a modest effect of dicarboxylic acids on uptake (see Section 2.1) and further confirm the strong influence of particle phase, liquid water content, and particulate nitrate concentration on N_2O_5 uptake rate.

Bertram and Thornton (2009) also proposed a mechanistic parameterization of $\gamma_{N_2O_5}$ that includes not only the competing effects of particle liquid water and nitrate but also chloride concentration,

$$\gamma_{N_2O_5} = A k' \left(1 - \frac{1}{\left(\frac{k_b[\text{H}_2\text{O}]}{k_a[\text{NO}_3^-]} \right) + 1 + \left(\frac{k_c[\text{Cl}^-]}{k_a[\text{NO}_3^-]} \right)} \right), \quad [11]$$

where A ($3.2 \times 10^{-8} \text{ s}$) is a constant that encompasses S , $c_{N_2O_5}$, and K_H , and k' can be calculated by using

$$k' = \beta - \beta e^{-\delta[\text{H}_2\text{O}]}, \quad [12]$$

with $\beta = 1.15 \times 10^6 \text{ s}^{-1}$ and $\delta = 1.3 \times 10^{-1} \text{ M}^{-1}$. The ratios of k_b to k_a and k_c to k_a are 6.0×10^{-2} and 29, respectively. Their parameterization is based on measurements on mixed chloride–nitrate particles, and the authors found that the presence of chloride can offset the suppression of $\gamma_{N_2O_5}$ due to nitrate. For aerosol that does not contain chloride, the parameterization by Bertram and Thornton (2009) is similar to the uptake reaction rate constant recommended by Griffiths et al. (2009), and their experimental results confirmed the dependence of $\gamma_{N_2O_5}$ on competing particle water content

and nitrate. Investigations regarding the impact of these chemistry parameterizations on the regional or global scale are still in progress. All the major efforts to parameterize N_2O_5 uptake thus far are summarized in Table 3.

Regarding the dependence of $\gamma_{N_2O_5}$ on aerosol nitrate content, the various parameterizations compare as shown in Figure 15 (bottom). All parameterizations show a decrease of $\gamma_{N_2O_5}$ as the nitrate content increases. As described in Section 2.1, the decrease of $\gamma_{N_2O_5}$ is due to the fact that the recombination reaction $\text{NO}_2^+ + \text{NO}_3^-$ is favored when large nitrate concentrations are present. The RH-independent parameterization for N_2O_5 uptake inhibition by the nitrate effect recommended by Riemer et al. (2003) corresponds mostly closely with other RH-dependent parameterization at 50% RH (Davis et al. 2008; Griffiths et al. 2009; Bertram and Thornton 2009). The greatest variations in $\gamma_{N_2O_5}$ by increasing nitrate content in the particle are found in the mid-RH range. Bertram and Thornton (2009) also recommended the need for including chloride ion, and the parameterized function indicates that the presence of the halogen species offsets the influence of the nitrate effect (line labeled "10% Cl⁻").

The connection between N_2O_5 chemistry and the chlorine cycle has been investigated by Simon et al. (2009), using the Comprehensive Air quality Model with extensions (CAMx). They estimated the effects of observed nitryl chloride concentrations on local chemistry in southeast Texas and found that ClNO_2 increases the total reactive chlorine mass by 20%–40% in this region; despite the high reactivity of chlorine, ClNO_2 caused only modest increases in ozone concentrations.

There have also been models that treat the heterogeneous interaction between N_2O_5 and water in the troposphere as one involving only cloud water. Leitch et al. (1988) studied the dissolution of N_2O_5 in cloud droplets by using a Lagrangian chemical model with winter conditions in northeastern region of the United States and compared the resulting NO_3^- content in cloud water with measured values. They showed that the mean production rate of HNO_3 via cloud scavenging is also much more efficient than homogeneous N_2O_5 hydrolysis, which was observed by Colville et al. (1994) as well. The potential significance of N_2O_5 uptake on fog droplets was demonstrated through measurements by Wood et al. (2005), where the N_2O_5 mixing ratio was well below the detection limit of the instrument (LIF) during nights of high RH (above 70%). Osthoff et al. (2006) encountered similar conditions, where N_2O_5 mixing ratio falls below the detection limit in the presences of fog. Using $\gamma_{N_2O_5} = 0.04$ on water droplet at 282 K (Vandoren et al. 1990), Sommariva et al. (2009) demonstrated that direct heterogeneous uptake of N_2O_5 by fog, when present, is the dominant atmospheric N_2O_5 loss pathway. The current IUPAC recommendation for heterogeneous N_2O_5 hydrolysis on water droplet has a negative temperature dependence,

$$\gamma_{N_2O_5} = 2.7 \times 10^{-5} \exp(1800/T). \quad [13]$$

TABLE 3
Summary of major N₂O₅ heterogeneous reaction probability ($\gamma_{\text{N}_2\text{O}_5}$) parameterization studies

Reference	Characteristics	Comments
Dentener and Crutzen (1993) Rierner et al. (2003)	Constant $\gamma_{\text{N}_2\text{O}_5} = 0.1$. Mixed sulfate and nitrate parameterization.	Generally over predicts N ₂ O ₅ uptake. Corresponds well with other RH and temperature-dependent parameterizations at 50% RH and 298 K.
Evans and Jacob (2005)	Categorized $\gamma_{\text{N}_2\text{O}_5}$ based upon aerosol composition: sulfate, organic carbon, black carbon, sea salt, and dust. $\gamma_{\text{N}_2\text{O}_5}$ for sulfate, organic carbon, and sea salt is functions of temperature and RH.	A wide range of aerosol types are considered.
Anttila et al. (2006)	Inhibition of aerosol uptake by organic coating is treated.	First to treat impact of organic coatings as a function of coating thickness and temperature.
Davis et al. (2008)	Rigorous statistical parameterization of $\gamma_{\text{N}_2\text{O}_5}$ as a function of the aerosol composition, RH, temperature, and phase state.	For inorganic aerosol, a wide range of temperature and RH are considered. Also, the phase of the particles is accounted.
Rierner et al. (2009)	Resistor model incorporating both mixed sulfate and nitrate parameterization (Rierner et al. 2003) and inhibition of aerosol uptake by organic coating (Anttila et al. 2006).	Comprehensive treatment of both inorganic species and organics. Nitrate and organics suppress $\gamma_{\text{N}_2\text{O}_5}$. Temperature and RH dependencies are not included.
Griffiths et al. (2009)	$\gamma_{\text{N}_2\text{O}_5}$ as a function of aerosol water and nitrate content.	On the basis of dicarboxylic-acids-containing sulfate aerosol experiments.
Bertram and Thornton (2009)	$\gamma_{\text{N}_2\text{O}_5}$ as a function of aerosol water, nitrate, and chloride content.	On the basis of measurements on mixed chloride–nitrate particles.

5.2. Modeled Tropospheric N₂O₅ Profile Response to Other Trace Gases and Meteorological Conditions

As described in Section 4.2, the N₂O₅ vertical distribution tends to increase with altitude and often reaches a maximum near or at the top of the NBL. These profiles are the result of a complex interplay between chemical production and loss processes combined with the vertical transport of pollutants. Resolving these profiles adequately poses an important modeling issue, because pollutants in the NBL may not be well mixed, especially in rural areas (Fish et al. 1999). If the vertical model resolution is too coarse, the results may overestimate N₂O₅ levels near the surface and underestimate above the NBL (Galmarini et al. 1997). One must account for not only the chemical processes among trace gases but also prevailing meteorological conditions that govern the atmospheric stability. To quantify the altitude dependence of tropospheric N₂O₅ profiles, several investigators implemented gas-phase and heterogeneous reactions as well as meteorological conditions in 1-D models (Rierner et al. 2003, 2009; Geyer and Stutz 2004).

Geyer and Stutz (2004) showed that the key factors controlling the N₂O₅ vertical gradient are surface NO emission strength and vertical gradients of VOC and particle surface area. Near the ground, when the simulated NO emissions are high,

available NO₃ at low altitudes undergoes titration and leads to a large indirect loss of N₂O₅. Vertical profiles of NO₃ are inversely related to the vertical gradient of VOC; the reaction between the two species reduces N₂O₅ levels in layers of the troposphere with high VOC concentrations. Depletion of NO₃ by NO and VOC, however, is balanced by the production of NO₃ via the reaction between NO₂ and O₃ (Rierner et al. 2003). N₂O₅ vertical profiles are positively correlated with NO₂, while negatively correlated with NO and VOC gradients. Figure 16 shows an example of calculated gas-phase vertical profiles for selected gas-phase species involved in N₂O₅ chemistry and illustrates the sensitivity of these profiles to atmospheric stability. The aerosol surface area gradient controls N₂O₅ levels directly by uptake through hydrolysis. While the reaction with NO is the dominant N₂O₅ loss pathway near the surface, at higher elevations, 20–80 m, the presence of high aerosol surfaces reduces the peak of N₂O₅ vertical profile in the NBL (Geyer and Stutz 2004). The reduced gradient in the N₂O₅ vertical profile in turn reduces the vertical transport of N₂O₅ to the surface layer. The role of aerosol surfaces on N₂O₅ profile is further complicated by the presence of nitrates and organic coatings on the aerosol that suppresses and inhibits the uptake of N₂O₅ on aerosols (Section 5.1).

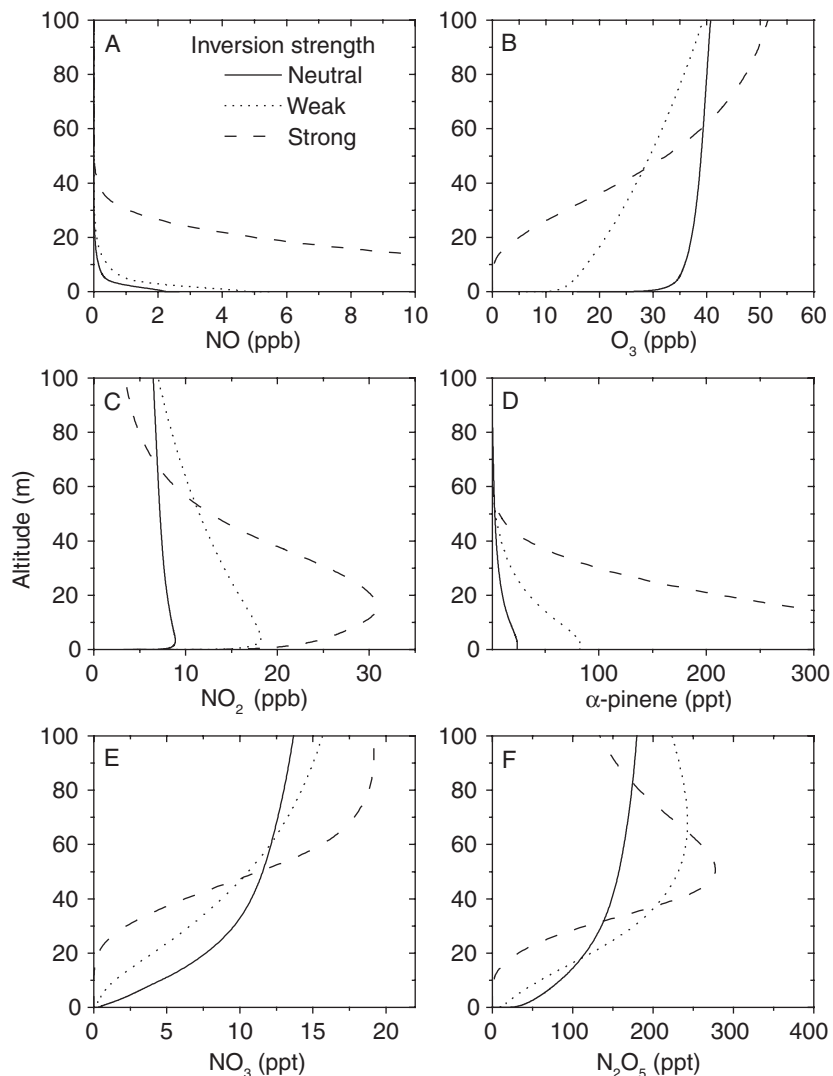


FIG. 16. Modeled NO, NO₂, NO₃, O₃, α -pinene, and N₂O₅ vertical profile responses to atmosphere stability (From Geyer and Stutz 2004).

Finally, numerical methods also present a source of uncertainty in simulating ambient N₂O₅ levels and spatial distributions. Though operator splitting in 3-D models is adequate for most tropospheric gaseous pollutants, it can be inadequate for fast-reacting species such as N₂O₅ (Nguyen and Dabdub 2003). Nguyen and Dabdub (2003) attributed N₂O₅ errors of as much as 44% to operator splitting. Nonoperator splitting methods avoid these errors, but their computational expense makes them impractical in most air quality model applications.

6. SUMMARY AND FUTURE RESEARCH NEEDS

Our understanding of the dynamics of N₂O₅ in the troposphere has improved significantly over the past few decades due to advances in laboratory measurements, ambient measurement techniques, and model developments. Although this review is not comprehensive with respect to the entire body of literature

on these areas, it provides (i) a review of the current understanding of the chemical mechanisms involving N₂O₅ formation and loss, along with its implications for atmospheric chemistry, (ii) a summary of the state-of-the-art measurement techniques available for making ambient in situ observations of N₂O₅ and the insights these measurements were able to provide, (iii) an examination of the uncertainties within the various methods of parameterization for multiphase/heterogeneous reactions of N₂O₅, and (iv) an identification of areas in need of further exploration and research.

With respect to the last point, the following is a list of the outstanding questions that could be addressed by further laboratory, field, and model studies in the near future.

Laboratory Experiments

- In order to obtain a more comprehensive parameterization of $\gamma_{\text{N}_2\text{O}_5}$, measurements of the heterogeneous

reaction probability under a wide range of relevant atmospheric conditions are needed. More specifically, the available literature lacks laboratory studies under conditions with low temperature and high RH.

- Detailed analysis of particles with mixed ice, sulfates, nitrates, chloride, and potentially other halide species requires closer examination.
- More research is needed on the distribution of the inorganic and organic phases in real atmospheric aerosol, along with a better-defined model parameterization of the potential effect of phase separation on N₂O₅ uptake coefficients.
- Further laboratory investigation of the third-order branch of the gas-phase reaction of N₂O₅ with water is required to resolve current discrepancies between laboratory and field determinations of the homogeneous reaction of N₂O₅ with water vapor.

Ambient Measurements

- Detailed understanding of the relative importance of N₂O₅ heterogeneous reaction on aerosols relative to other losses (e.g., indirect, oxidative NO₃ loss) should be determined from a larger body of field data that includes simultaneous measurements of N₂O₅, its precursors, VOC, and the physical and chemical properties of the aerosol, as well as meteorological parameters.
- The role of dry deposition as a loss process for N₂O₅ is poorly understood, especially in shallow surface layers and/or the nocturnal boundary layer. Further field studies examining vertical gradients in the lowest layers of the nocturnal boundary layer are needed.
- N₂O₅ chemistry is most prevalent in colder and darker periods of the year, where currently there is a smaller focus based on recent field studies. Understanding how N₂O₅ behaves in the winter season will be important to issues such as nitrate aerosol formation, global NO_x lifetimes and oxidant burdens, and halogen activation.
- Vertical profiles (e.g., from ground-based DOAS at various levels or from tall towers) and aircraft measurements of N₂O₅, related trace gases and aerosol within and above the NBL, and the overlying residual daytime boundary layer are needed to understand interactions between surface-level emissions, mixing, and chemistry that occurs aloft in the nighttime atmosphere.
- The existing database for simultaneous measurement of ClNO₂ and N₂O₅ remains too limited to accurately quantify the efficiency and extent of nighttime halogen activation.

Model Development

- There is the need for chemical transport models with high vertical resolutions that quantify the vertical transport of various trace gases and high horizontal resolutions that account for N₂O₅ production and destruction as a nonlinear function of NO_x concentration and sur-

face area density. Model studies require vertical resolutions below 10 m in the nocturnal boundary layer, in particular near the ground, and several layers between 10 and 100 m. The description of vertical mixing in stable boundary layers remains a challenge and requires more research.

- In addition to the aerosol surface, the urban canopy can also provide the surface area for heterogeneous and surface reactions to occur in the urban environment. Chemical transport models currently do not consider this contribution.
- The more detailed the parameterizations become for $\gamma_{\text{N}_2\text{O}_5}$, the more assumptions need to be made regarding the aerosol mixing state, i.e., which fraction of the populations contains nitrate/chloride/organic coatings. This is a challenge for the representation of aerosol particles in chemical transport models.
- There is evidence from field measurements that the N₂O₅ hydrolysis on fog and cloud droplets could have a significant impact on the N₂O₅ budget. Modeling studies on this topic are rare but would help further our understanding in this regard, as measurement techniques have difficulties operating in cloudy or foggy environments.
- Model studies on ambient N₂O₅ profiles in the free troposphere, upper troposphere, and lower stratosphere are currently lacking. The sources of NO_x in this region are a major source of uncertainty, in particular the contribution of lightning-generated NO_x and the effect of its vertical distribution on the formation of N₂O₅.

Nomenclature

Acronyms used in this article

BBCEAS	BroadBand Cavity Enhanced Absorption Spectroscopy
CEAS	Cavity-Enhanced Absorption Spectroscopy
CIMS	Chemical Ionization Mass Spectrometry
CRDS	Cavity Ring-Down Spectroscopy
cw	Continuous Wave
DMS	Dimethyl Sulfide
DOAS	Differential Optical Absorption Spectroscopy
DOY	Day of Year
GEOS	Goddard Earth Observing System
ICARTT	International Consortium for Atmospheric Research on Transport and Transformation
ICEALOT	International Chemistry Experiment in the Arctic Lower Troposphere
ID	Ion Drift
IMR	Ion-Molecule Reaction Region
ITCT	Intercontinental Transport and Chemical Transformation
LED	Light Emitting Diode
LIF	Laser-Induced Fluorescence
MBL	Marine Boundary Layer

NBL	Nighttime Boundary Layer
NEAQS	New England Air Quality Study
NOAA	National Oceanic and Atmospheric Administration
OA	Off-axis
PAN	Peroxy Acetyl Nitrate
PMCAMx	Particulate Matter Comprehensive Air Quality Model with eXtensions
PSCRDS	Phase Shift Cavity Ring-Down Spectroscopy
RH	Relative Humidity
RHaMBLe	Reactive Halogens in the Marine Boundary Layer
SAPHIR TD	Simulation of Atmospheric PHotochemistry In a large Reaction Chamber
TD	Thermal Dissociation
TexAQS	Texas Air Quality Study
UTC	Universal Time Coordinated
UW	University of Washington
VOC	Volatile Organic Compound

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